# 2013 Update of the South African Hake Reference Case Assessment 

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## Summary


#### Abstract

The Reference Case assessment of the SA hake resource is updated using unchanged methodology from that used in developing the existing OMP, but with revised and updated data sets. The only change of particular note is the inclusion of new longline catch-at-length data subsequent to 2000 results in a somewhat reduced estimate of current the $M$. paradoxus spawning biomass.


## INTRODUCTION

This paper presents an update of the South African hake Reference Case assessment (RS1) (Rademeyer and Butterworth, 2010). Compared to the 2012 routine update of this assessment (Rademeyer, 2012), it includes new commercial (catches, length distribution and CPUE) data, but no new survey data are available.

## METHODOLOGY and DATA

The updated data are listed in Appendix A. The methodology is detailed in Appendix B.
Six assessments are compared:

1) RS1-2012: the 2012 routine update presented last year (Rademeyer, 2012);
2) RS1-2013a: the comparable assessment to RS1-2012 with updated catches and CPUE data to 2012;
3) RS1-2013b: as RS1-2013a, with updated maturity-at-length (Singh et al. 2011) and length-weight parameters (Singh 2013);
4) RS1-2013c: as RS1-2013b, with further commercial offshore trawl CAL data (see Appendix 1 for details);
5) RS1-2013d: as RS1-2013c, with further commercial longline CAL data (see Appendix 1 for details);
6) RS1-2013e: as RS1-2013d, with different selectivity curves estimated for the longline fleet on the West and South Coasts for M. capensis (in RS1-2013d, the longline selectivity curve for M. capensis is taken to be the same on the West and South Coasts) and with three periods with different selectivity curves for the longline fleet: i) pre-2000, ii) 2000-2005 and iii) 2006 to current. The changes in selectivity are reflected by shifts of the ascending limb of the logistic curve, separately for each species and coast, while the steepness of the limb and the slope at older ages are kept the same (i.e. seven further parameters estimated compared to RS1-2013d). The choice of the years for the changes in selectivity was made by inspection of plots of residuals. This assessment is termed the "new Reference Case" (new RC);

## RESULTS and DISCUSSION

## From the 2012 assessment to the 2013 new RC

Table 1 compares estimates of management quantities for the six assessments, while Fig. 1 plots the spawning biomass trajectories. Fig. 2 compares the recruitment trajectories and Fig. 3 plots the estimated stock-recruitment relationships.

Figs 4 and 5 compare the fits to the CPUE and survey abundance indices for RS1-2012 and the new RC. The fits are reasonable for all series, but there are indications of correlation in residuals over time.

## The 2013 new RC

The estimated commercial and survey selectivity curves for the new $R C$ are shown in Figs $6 a$ and $6 b$ respectively. Apart from $M$. paradoxus on the South Coast, the male and female selectivities-at-length are assumed to be the same and are then converted to gender-specific selectivities-at-age. Because of selectivity differences between males and females evident for the South Coast surveys for M. paradoxus, gender-specific selectivities are estimated for this species in the South Coast autumn and spring surveys, with the female selectivities (for M. paradoxus only) scaled downwards for these two surveys by a common factor across lengths which is estimated in the model fitting procedure. This gender difference is assumed to effect the commercial fleet as well, and the female selectivity for the South Coast offshore trawl fleet (the only fleet assumed to catch M. paradoxus on the South Coast) is therefore also scaled downwards by a factor estimated in the model fitting. The female selectivity scaling factors estimated for the South Coast spring and autumn surveys are 0.42 and 0.90 respectively. The scaling factor for the offshore trawl fleet is taken to be the average of these two values.

All the commercial selectivity curves show a decrease for large fish, apart for the South Coast offshore trawl fleet for M. paradoxus. This decrease is estimated for all fleets apart from the offshore trawl and South Coast handline fleets for $M$. capensis. For the $M$. capensis offshore trawl fleet, the selectivity slope (the rate of decrease of selectivity at larger sizes) is fixed to $1 / 3$ of that estimated for the inshore trawl fleet, while for the handline fleet, this slope is taken to be the average of the estimated South Coast longline and inshore fleet slopes. (It should be noted that the RC makes particular assumptions about the values of the natural mortality-at-age schedule, specifically the natural mortality is fixed at 0.75 for ages 0 and 1 and fixed at 0.375 for ages 6 and above, with a linear trend between these two values for ages 2 to 5 . Selectivities will change if these assumptions are changed; results for such changes will be reported in a following paper.)

The fits to the survey gender-aggregated and gender-disaggregated catch-at-length data are shown in Figs 7 and 8 respectively. These fits are broadly reasonable.

The fits to the commercial catch-at-length data are shown in Fig. 9. The fits are averaged over all the years for which data are available, and are reasonable for all the data sets. There are however some patterns evident in the bubble plots of residuals which could perhaps be improved by adding further periods between which selectivities change.

Figs 10 and 11 plot the gender-specific growth curves and length-at-age distributions estimated in the new RC. The difference between male and female growth curves is estimated to be greater for M. paradoxus than for M. capensis. $M$. capensis is estimated to grow to slightly larger sizes than $M$. paradoxus.

The fits to the ALKs are given in Appendix B.

## Including the new (2000 onwards) longline catch-at-length data

Although results do change with the inclusion of the updated catches, CPUE (RS1-2013a), with the change in the maturity-at-length and length weight relationships (RS1-2013b) and with the inclusion of the new offshore trawl catches (RS1-2013c), the magnitude of the changes is not large. The current status of the resource $s$ is broadly
unchanged, with M. paradoxus close to MSY level and M. capensis well above it. The picture however is rather different for M. paradoxus when the new longline CAL data are included (RS1-2013d and RS1-2013e).

When these new longline data are included, it is evident that a change in selectivity over time is necessary for the model to fit to the data (Fig. 12). Two changes have been made to the manner in which longline selectivities are modelled compared to preceding analyses in moving to RS1-2013e. The first is that with the availability of speciesdisaggregated longline data for the West Coast, a selectivity curve can be estimated directly for M. capensis, rather than assuming the same selectivity curve as on the South Coast. Secondly, three periods of differing selectivities have been assumed to better reflect the patterns in these CAL data, namely pre-2000, 2000-2005 and 2006 onwards. The differing-selectivity periods are taken to apply to both $M$. paradoxus and $M$. capensis. The changes in selectivity curves are reflected by estimating a shift of the ascending limb of the logistic curve separately for each species and coast. The steepness of the limb and the slope at older ages are kept the same (i.e. seven further parameters estimated compared to RS1-2013d). The choice of the years for the changes in selectivity was made by inspecting plots of residuals. These changes are largely justified in AIC terms with an improvement of over 19 likelihood points. Fig. 12 compares the residuals for the fit to the longline proportion-at-length with and without these changes. The patterns in the residuals are much reduced in the new RC compared to RS1-2013d. Note that the trend in the pattern of these selectivity curves (see Fig. 6a) is compatible with general perceptions - that over time this fishery has shifted its focus to include more of the smaller hake.

The question remains as to why including new longline CAL data results in a rather different perception of the current status of $M$. paradoxus. To investigate this further, the new RC was also run downweighting the new longline CAL data in the negative log-likelihood (with $w=0.0001$ instead of $w=0.1$ - recall that CAL data are routinely downweighted by this $10 \%$ multiplicative factor for reasons detailed in Appendix $B$ - see text following equation B.37). The results are compared to the new RC results in Table 2 and the contributions to the total negative log-likelihood compared in Table 3. Fig. 14 shows that some of the CAL misfit in the downweighting case is corrected to a fair extent when these data are given full weight. Comparison of the negative log-likelihood contributions in Table 1 for RS1-2013c and RS12013e shows that the changed result is linked to some "tension" between the abundance indices and the commercial CAL data. In changing from the former to the latter assessment, the fits to all the abundance indices deteriorate, but improve for both the commercial trawl (slightly) and the longline CAL data. In adding further commercial CAL data, the net weight accorded to these data increases relative to the abundance index data, leading to the change in results.

## CONCLUDING REMARKS

Developing further commentary on the results of these assessments has been deferred in the interests of early circulation. The result of particular note is the somewhat worse status estimated for $M$. paradoxus when the further commercial longline CAL data are taken into account. This will be explored further in a subsequent paper examining various sensitivities to the new RC, which will also include estimates of precision for the results for the new RC.

## REFERENCES

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Table 1: Comparison of estimates of management quantities of the $M$. paradoxus and $M$. capensis coast-combined resources for the six assessments. MSY and associated quantities are given for the offshore trawl fleet. Biomass units are thousand tons. Note that not all the -InL values are comparable given that different data are used. $K^{s p}$, $B_{y}^{s p} / K^{s p}, B_{M S Y}^{s p} / K^{s p}$ and $B_{y}^{s p} / B_{M S Y}^{s p}$ are all in terms of the female component of the spawning biomass only.

|  | RS1-2012 | RS1-2013a | RS1-2013b | RS1-2013c | RS1-2013d | RS1-2013e |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| -InL total | -58.1 | -123.4 | -117.3 | -124.0 | -153.2 | -172.3 |
| CPUE historic | -38.1 | -40.2 | -40.3 | -40.2 | -39.8 | -39.9 |
| CPUE GLM | -143.0 | -170.3 | -172.2 | -170.2 | -169.5 | -168.6 |
| Survey | -38.0 | -33.8 | -33.3 | -33.4 | -31.3 | -32.1 |
| Commercial CAL - trawl | -35.7 | -44.1 | -35.1 | -44.9 | -42.9 | -45.1 |
| Commercial CAL - longline | -16.2 | -16.0 | -16.1 | -16.1 | -50.5 | -68.9 |
| Survey CAL (sex-aggr.) | -2.3 | -5.2 | -4.6 | -5.2 | -2.9 | -3.0 |
| Survey CAL (sex-disaggr.) | 66.7 | 43.1 | 42.7 | 43.0 | 43.0 | 42.9 |
| ALK | 124.0 | 118.2 | 118.2 | 118.4 | 116.3 | 117.4 |
| Recruitment penalty | 8.8 | 8.7 | 7.6 | 8.4 | 8.9 | 8.6 |
| Selectivity smoothing penalty | 15.5 | 16.0 | 15.5 | 16.0 | 15.4 | 15.5 |
| $K^{\text {sp }}$ | 586 | 834 | 875 | 859 | 729 | 754 |
| $h$ | 1.23 | 0.93 | 0.92 | 0.93 | 0.97 | 0.96 |
| $B^{\text {Sp }}{ }_{2012}$ | 134 | 161 | 149 | 165 | 104 | 114 |
| $\simeq B^{\text {Sp }}{ }_{2012} / K^{\text {Sp }}$ | 0.23 | 0.19 | 0.17 | 0.19 | 0.14 | 0.15 |
| ${ }_{0}{ }_{0} B^{\text {Sp }}{ }_{2013}$ | - | 158 | 147 | 162 | 98 | 108 |
| O $B^{\text {Sp }} 2013 / K^{\text {sp }}$ | - | 0.19 | 0.17 | 0.19 | 0.13 | 0.14 |
| ${ }_{2} B^{\text {SP }}{ }_{M S \gamma}$ | 134 | 164 | 172 | 174 | 152 | 158 |
| $\sum B^{s p}{ }_{M S Y} / K^{\text {Sp }}$ | 0.23 | 0.20 | 0.20 | 0.20 | 0.21 | 0.21 |
| $B^{5 p}{ }_{2012} / B^{\text {Sp }}{ }_{M S Y}$ | 0.98 | 0.98 | 0.87 | 0.95 | 0.69 | 0.72 |
| $B^{5 p}{ }_{2013} / B^{\text {sp }}{ }_{M S Y}$ | - | 0.96 | 0.85 | 0.93 | 0.65 | 0.68 |
| MSY | 113 | 116 | 117 | 116 | 111 | 113 |
| $K^{\text {sp }}$ | 251 | 288 | 244 | 244 | 240 | 239 |
| $h$ | 1.40 | 1.02 | 0.99 | 0.99 | 1.04 | 1.03 |
| $B^{s p}{ }_{2012}$ | 240 | 186 | 156 | 154 | 151 | 152 |
| $\cong B^{5 p}{ }_{2012} / K^{5 p}$ | 0.96 | 0.65 | 0.64 | 0.63 | 0.63 | 0.64 |
| $\frac{\tilde{n}}{\hat{y}} \quad B^{s p}{ }_{2013}$ | - | 207 | 174 | 172 | 170 | 170 |
| O $B^{\text {Sp }}{ }_{2013} / K^{\text {sp }}$ | - | 0.72 | 0.71 | 0.71 | 0.71 | 0.71 |
| $\pm B^{s p}{ }_{M S Y}$ | 89 | 109 | 100 | 100 | 95 | 96 |
| ${ }^{*} B^{\text {sp }}{ }_{M S Y} / K^{\text {sp }}$ | 0.36 | 0.38 | 0.41 | 0.41 | 0.39 | 0.40 |
| $B^{5 p}{ }_{2012} / B^{\text {sp }}{ }_{M S Y}$ | 2.00 | 1.71 | 1.56 | 1.54 | 1.60 | 1.58 |
| $B^{5 p}{ }_{2013} / B^{\text {sp }}{ }_{M S Y}$ | - | 1.91 | 1.74 | 1.72 | 1.80 | 1.76 |
| MSY | 70 | 62 | 63 | 64 | 63 | 63 |

Table 2: Comparison of estimates of management quantities of the $M$. paradoxus and $M$. capensis coast-combined resources for the new RC (RS1-2013e) and the run downweighting the new longline CAL data. MSY and associated quantities are given for the offshore trawl fleet. Biomass units are thousand tons. Note that the -InL values are not comparable given that different data are used. $K^{s p}, B_{y}^{s p} / K^{s p}, B_{M S Y}^{s p} / K^{s p}$ and $B_{y}^{s p} / B_{M S Y}^{s p}$ are all in terms of the female component of the spawning biomass only.

|  |  | RS1-2013e | Downweighting new longline CAL |
| :---: | :---: | :---: | :---: |
|  | -InL total | -172.3 | -125.3 |
|  | CPUE historic | -39.9 | -40.2 |
|  | CPUE GLM | -168.6 | -170.2 |
|  | Survey | -32.1 | -33.4 |
|  | Commercial CAL - trawl | -45.1 | -44.9 |
|  | Commercial CAL - longline | -68.9 | -17.4 |
|  | Survey CAL (sex-aggr.) | -3.0 | -5.3 |
|  | Survey CAL (sex-disaggr.) | 42.9 | 43.1 |
|  | ALK | 117.4 | 118.4 |
|  | Recruitment penalty | 8.6 | 8.5 |
|  | ectivity smoothing penalty | 15.5 | 15.9 |
| $\begin{aligned} & \text { n } \\ & \text { x } \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & \text { a } \\ & \dot{\Sigma} \end{aligned}$ | $K^{\text {Sp }}$ | 754 | 854 |
|  | $h$ | 0.96 | 0.93 |
|  | $B^{\text {sp }} 2012$ | 114 | 163 |
|  | $B^{5 p}{ }_{2012} / K^{\text {sp }}$ | 0.15 | 0.19 |
|  | $B^{5 p}{ }_{2013}$ | 108 | 160 |
|  | $B^{\text {sp }}{ }_{2013} / K^{\text {sp }}$ | 0.14 | 0.19 |
|  | $B^{s p}{ }_{M S Y}$ | 158 | 173 |
|  | $B^{s p}{ }_{M S Y} / K^{s p}$ | 0.21 | 0.20 |
|  | $B^{\text {sp }}{ }_{2012} / B^{\text {sp }}{ }_{M S Y}$ | 0.72 | 0.94 |
|  | $B^{\text {Sp }}{ }_{2013} / B^{\text {sp }}{ }_{M S Y}$ | 0.68 | 0.92 |
|  | MSY | 113 | 116 |
| $\begin{aligned} & \tilde{W} \\ & \tilde{U} \\ & 0 \\ & 0 \\ & \dot{8} \end{aligned}$ | $K^{\text {Sp }}$ | 239 | 243 |
|  | $h$ | 1.03 | 0.99 |
|  | $B^{\text {Sp }}{ }_{2012}$ | 152 | 154 |
|  | $B^{\text {Sp }}{ }_{2012} / K^{\text {SP }}$ | 0.64 | 0.63 |
|  | $B^{5 p}{ }_{2013}$ | 170 | 172 |
|  | $B^{5 p}{ }_{2013} / K^{5 p}$ | 0.71 | 0.71 |
|  | $B^{s p}{ }_{M S Y}$ | 96 | 100 |
|  | $B^{s p}{ }_{M S \gamma} / K^{s p}$ | 0.40 | 0.41 |
|  | $B^{\text {Sp }}{ }_{2012} / B^{\text {sp }}{ }_{M S Y}$ | 1.58 | 1.54 |
|  | $B^{5 p}{ }_{2013} / B^{\text {sp }}{ }_{M S Y}$ | 1.76 | 1.72 |
|  | MSY | 63 | 63 |

Table 3: Contribution to the negative log-likelihood for the new RC (RS1-2013e) and the assessment downweighting the new longline CAL data.

|  |  | spp combined | RS1-201 | cap | Downwe spp combined | hting ne <br> para | LL CAL <br> cap |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| -InL total | WC | -172.30 |  |  | -125.26 |  |  |
| CPUE historic | WC | -29.61 |  |  | -29.74 |  |  |
|  | SC | -10.31 |  |  | -10.44 |  |  |
| CPUE GLM | WC |  | -49.83 | -38.21 |  | -47.04 | -39.51 |
|  | SC |  | -42.65 | -37.86 |  | -46.59 | -37.09 |
| Survey | WC summer |  | -11.21 | -3.94 |  | -12.00 | -4.03 |
|  | WC winter |  | -3.22 | 1.01 |  | -3.33 | 1.02 |
|  | SC autumn |  | 1.69 | -7.49 |  | 1.98 | -7.56 |
|  | SC spring |  | 6.72 | -15.62 |  | 6.31 | -15.78 |
| Commercial CAL | WC offshore | -22.52 |  |  | -22.92 |  |  |
|  | SC offshore | 2.24 |  |  | 2.37 |  |  |
|  | BC offshore | -3.19 |  |  | -3.39 |  |  |
|  | SC inshore |  |  | -21.61 |  |  | -20.98 |
|  | WC longline (1994-1997) | -11.08 |  |  | -11.27 |  |  |
|  | WC longline ( $2000+$ ) |  | -0.24 | -0.13 |  | -0.16 | -0.09 |
|  | SC longline (1994-1997) |  |  | -4.79 |  |  | -5.75 |
|  | SC longline (2000+) |  |  | -0.16 |  |  | -0.10 |
| Survey CAL |  |  |  |  |  |  |  |
| Sex-aggregated | WC summer |  | -6.89 | 11.65 |  | -7.40 | 11.38 |
|  | WC winter |  | -2.96 | 6.02 |  | -3.07 | 5.87 |
|  | SC autumn |  | 2.64 | -6.31 |  | 2.53 | -7.04 |
|  | SC spring |  | 3.69 | -10.81 |  | 3.62 | -11.24 |
| Sex-disaggregated | WC summer |  | -0.41 | 51.57 |  | -1.07 | 51.07 |
|  | WC winter |  | - | - |  | - | - |
|  | SC autumn |  | 3.07 | -5.66 |  | 3.09 | -5.31 |
|  | SC spring |  | 17.64 | -23.26 |  | 18.43 | -23.11 |
| ALK |  | 117.41 |  |  | 118.41 |  |  |
| Recruitment penalty |  |  | 3.65 | 4.96 |  | 3.40 | 5.06 |
| Sel. smoothing penalty |  | 15.47 |  |  | 15.95 |  |  |



Fig. 1a: Trajectories of female spawning biomass (in terms of its pre-exploitation level) for the five assessments.The horizontal lines represent MSYL.


Fig. 1b: As above but with different scales on both axes to focus on more recent trends.
M. paradoxus

M. capensis


Fig. 2: Time series of recruitment for the new RC (RS1-2013e) compared to RS1-2012. Note that the decrease in $\sigma_{R}$ from 0.25 to 0.1 has been moved one year forward for the new RC compared to RS1-2012.


Fig. 3: Estimated stock-recruitment relationships for RS1-2012 (dashed blue line and blue crosses) and the new RC (RS1-2013e) (solid black line and black dots).


Fig. 4: Fits to the CPUE abundance indices for the RS1-2012 (dashed blue line) and the new RC (solid black line).


Fig. 5: Fits to the west coast summer and south coast autumn abundance series from surveys by Africana (the two longest series) for the RS1-2012 (dashed blue line) and the new RC (solid black line) assessments. The observed values shown as $\Delta$ were conducted by the Africana with the new gear and have been rescaled by the agreed calibration factor for the species concerned.

Selectivity-at-length (gender independent)


Fig. 6a: Commercial selectivities-at-length estimated for the new RC. For the offshore trawl fleet, the selectivity periods are as follows: i) first period: 1917-1976, ii) second period: 1977-1984 and iii) third period: 1993-2013. A linear change is taken between 1984 and 1993.


Fig. 6b: Survey selectivities-at-length estimated for the new RC.

## M. paradoxus

## M. capensis



 are grey, whereas for negative residuals the bubbles are white.

FISHERIES/2013/NOV/SWG-DEM67
M. paradoxus







M. capensis








Fig. 8a: Fit of RS1-2013e to the west coast summer survey gender-disaggregated proportion-at-length data (in some plots, aggregated over years for which data are available).
M. paradoxus













Fig. 8b: Fit of RS1-2013e to the south coast spring survey gender-disaggregated proportion-at-length data (in some plots, aggregated over years for which data are available).

FISHERIES/2013/NOV/SWG-DEM67
M. paradoxus







MARAM IWS/DEC13/Hake/P2
M. capensis

## 








Fig. 8c: Fit of RS1-2013e to the south coast autumn survey gender-disaggregated proportion-at-length data (in some plots, aggregated over years for which data are available).


Fig. 9a: Fit of RS1-2013e to the commercial offshore and inshore trawl proportion-at-length data, aggregated over years for which data are available. Bubble plots of the corresponding residuals are shown.

FISHERIES/2013/NOV/SWG-DEM67





MARAM IWS/DEC13/Hake/P2


| 2010 <br> 2009 <br> 2008 <br> 2007 <br> 2006 <br> 2005 <br> 2004 <br> 2003 <br> 2002 <br> 2001 <br> 2000$-$ |  |
| :---: | :---: |
| 35 | 45 55 Length $(\mathrm{cm})$ 75 85 |



Fig. 9b: Fit of RS1-2013e to the commercial longline proportion-at-length data, aggregated over years for which data are available. Bubble plots of the corresponding residuals are shown.
M. paradoxus

M. capensis


Fig. 10: Estimated growth curves for the new RC (RS1-2013e).


Fig. 11: Estimated distribution of length-at-age for the new RC (RS1-2013e).


Fig. 12: Comparison of bubble plots of residuals for the fit to the longline proportion-at-length data with constant selectivity RS1-2013d, left-hand plots) and with a changes in selectivity in 1999/2000 and 2005/2006 (RS1-2013e, right-hand plots).


Fig. 13: Longline CAL data input to the assessment, averaged over the years indicated.

RS1-2013e: BC

downweighting new longline CAL

Fig. 14: Comparison of fit to the longline proportion-at-length data, aggregated over years for which data are available, for the new RC (RS1-2013e) and the assessment downweighting the new (2000+) CAL longline data.


Fig. 15: Comparison of model predicted (new RC) and observed longline CPUE (Somhlaba, pers. commn). These data are not fitted in the new RC as they are from a dated analysis (the GLM-standardisation of more recent data is under finalisation).

## Appendix A - Data Tables

Data that are shaded represent new or revised information since the 2012 assessment (Rademeyer, 2012).
Table App.A.1a: Species-disaggregated catches (in thousand tons) by fleet of South African hake from the south and west coasts for the period 1917-1977. The baseline assessment assumes 1958 as the centre year of the shift from a primarily M. capensis to a primarily M. paradoxus in the offshore trawl catches.

|  | M. paradoxus Offshore WC | M. capensis Offshore WC |
| :---: | :---: | :---: |
| 1917 | - | 1.000 |
| 1918 | - | 1.100 |
| 1919 | - | 1.900 |
| 1920 | - | 0.000 |
| 1921 | - | 1.300 |
| 1922 | - | 1.000 |
| 1923 | - | 2.500 |
| 1924 | - | 1.500 |
| 1925 | - | 1.900 |
| 1926 | - | 1.400 |
| 1927 | - | 0.800 |
| 1928 | - | 2.600 |
| 1929 | - | 3.800 |
| 1930 | - | 4.400 |
| 1931 | - | 2.800 |
| 1932 | - | 14.300 |
| 1933 | - | 11.100 |
| 1934 | - | 13.800 |
| 1935 | - | 15.000 |
| 1936 | - | 17.700 |
| 1937 | - | 20.200 |
| 1938 | - | 21.100 |
| 1939 | - | 20.000 |
| 1940 | - | 28.600 |
| 1941 | - | 30.600 |
| 1942 | 0.001 | 34.499 |
| 1943 | 0.001 | 37.899 |
| 1944 | 0.002 | 34.098 |
| 1945 | 0.004 | 29.196 |
| 1946 | 0.011 | 40.389 |
| 1947 | 0.021 | 41.379 |


|  | M. paradoxus |  | M. capensis |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Offshore |  | Offshore |  | Inshore |
|  | WC | SC | WC | SC | SC |
| 1948 | 0.059 | - | 58.741 | - | - |
| 1949 | 0.113 | - | 57.287 | - | - |
| 1950 | 0.275 | - | 71.725 | - | - |
| 1951 | 0.662 | - | 88.838 | - | - |
| 1952 | 1.268 | - | 87.532 | - | - |
| 1953 | 2.558 | - | 90.942 | - | - |
| 1954 | 5.438 | - | 99.962 | - | - |
| 1955 | 10.924 | - | 104.476 | - | - |
| 1956 | 19.581 | - | 98.619 | - | - |
| 1957 | 34.052 | - | 92.348 | - | - |
| 1958 | 51.895 | - | 78.805 | - | - |
| 1959 | 76.609 | - | 69.391 | - | - |
| 1960 | 100.490 | - | 59.410 | - | 1.000 |
| 1961 | 104.009 | - | 44.691 | - | 1.308 |
| 1962 | 109.596 | - | 38.004 | - | 1.615 |
| 1963 | 129.966 | - | 39.534 | - | 1.923 |
| 1964 | 126.567 | - | 35.733 | - | 2.231 |
| 1965 | 159.704 | - | 43.296 | - | 2.538 |
| 1966 | 154.109 | - | 40.891 | - | 2.846 |
| 1967 | 139.973 | 7.086 | 36.727 | 7.100 | 3.154 |
| 1968 | 113.890 | 13.958 | 29.710 | 13.950 | 3.462 |
| 1969 | 131.023 | 18.982 | 34.077 | 18.948 | 3.769 |
| 1970 | 113.124 | 11.876 | 29.376 | 11.847 | 4.077 |
| 1971 | 160.384 | 15.078 | 41.616 | 15.037 | 4.385 |
| 1972 | 193.694 | 23.382 | 50.239 | 23.314 | 4.692 |
| 1973 | 125.292 | 36.232 | 32.490 | 36.124 | 5.000 |
| 1974 | 97.674 | 45.496 | 25.326 | 45.357 | 10.056 |
| 1975 | 71.165 | 33.783 | 18.452 | 33.680 | 6.372 |
| 1976 | 114.268 | 26.005 | 29.626 | 25.925 | 5.740 |
| 1977 | 81.260 | 18.515 | 21.068 | 18.457 | 3.500 |

Table App.A.1b: Species-disaggregated catches (in thousand tons) by fleet of South African hake from the south and west coasts for the period 1978-present. The new/revised offshore trawl catches are from Glazer (2013) (the catches in Glazer, 2013, were subsequently updated to include 2012, Glazer, pers. commn), the new inshore and handline catches are from Rob Cooper (pers. comm.) and the new longline catches from Sobahle Somhlaba (pers. comm.). For 2013, the catches are taken as the 2013 TAC with the same proportion by species and fleet as in 2012.

|  | M. paradoxus |  |  | M. capensis |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Offshore |  | Longline WC | Offshore |  | InshoreSC | Longline |  | Handline SC |
|  | WC | SC |  | WC | SC |  | WC | SC |  |
| 1978 | 107.701 | 4.937 | - | 19.812 | 2.648 | 4.931 | - | - | - |
| 1979 | 101.890 | 3.575 | - | 31.633 | 3.345 | 6.093 | - | - | - |
| 1980 | 105.483 | 3.676 | - | 28.045 | 2.784 | 9.121 | - | - | - |
| 1981 | 95.330 | 1.767 | - | 25.601 | 3.719 | 9.400 | - | - | - |
| 1982 | 88.933 | 5.057 | - | 24.417 | 6.300 | 8.089 | - | - | - |
| 1983 | 74.173 | 7.034 | 0.161 | 20.260 | 5.482 | 7.672 | 0.069 | - | - |
| 1984 | 86.045 | 5.718 | 0.256 | 25.210 | 5.217 | 9.035 | 0.110 | 0.016 | - |
| 1985 | 98.283 | 12.694 | 0.817 | 26.788 | 7.322 | 9.203 | 0.350 | 0.292 | 0.065 |
| 1986 | 107.907 | 11.539 | 0.965 | 25.898 | 4.427 | 8.724 | 0.413 | 0.302 | 0.084 |
| 1987 | 96.162 | 10.536 | 2.500 | 21.363 | 5.148 | 8.607 | 1.071 | 0.353 | 0.096 |
| 1988 | 83.606 | 8.664 | 3.628 | 22.976 | 5.852 | 8.417 | 1.555 | 0.331 | 0.071 |
| 1989 | 85.298 | 9.039 | 0.203 | 21.961 | 9.873 | 10.038 | 0.087 | 0.032 | 0.137 |
| 1990 | 84.969 | 13.622 | 0.270 | 18.668 | 9.169 | 10.012 | 0.116 | - | 0.348 |
| 1991 | 89.371 | 15.955 | - | 17.079 | 6.119 | 8.206 | - | 3.000 | 1.270 |
| 1992 | 86.777 | 22.368 | - | 16.510 | 4.094 | 9.252 | - | 1.500 | 1.099 |
| 1993 | 105.114 | 12.472 | - | 12.951 | 1.789 | 8.870 | - | 0.000 | 0.278 |
| 1994 | 106.287 | 8.588 | 1.130 | 17.580 | 2.464 | 9.569 | 0.484 | 0.626 | 0.449 |
| 1995 | 102.877 | 5.395 | 0.670 | 18.020 | 1.755 | 10.630 | 0.287 | 0.650 | 0.756 |
| 1996 | 110.460 | 11.080 | 1.676 | 18.715 | 2.209 | 11.062 | 0.718 | 1.828 | 1.515 |
| 1997 | 103.035 | 13.651 | 1.806 | 14.119 | 2.185 | 8.834 | 0.774 | 1.872 | 1.404 |
| 1998 | 113.083 | 11.703 | 0.647 | 14.570 | 2.450 | 8.283 | 0.277 | 1.471 | 1.738 |
| 1999 | 89.147 | 13.435 | 1.963 | 14.614 | 1.912 | 8.595 | 0.841 | 4.144 | 2.749 |
| 2000 | 97.417 | 9.920 | 3.294 | 20.285 | 3.610 | 10.906 | 1.412 | 2.077 | 5.500 |
| 2001 | 101.990 | 11.016 | 2.656 | 15.606 | 5.141 | 11.836 | 1.138 | 1.688 | 7.300 |
| 2002 | 91.720 | 15.445 | 4.802 | 13.211 | 3.140 | 9.581 | 2.058 | 3.945 | 3.500 |
| 2003 | 95.143 | 21.107 | 4.081 | 10.233 | 3.926 | 9.883 | 1.749 | 4.878 | 3.000 |
| 2004 | 86.916 | 30.746 | 3.606 | 11.315 | 4.024 | 10.004 | 1.546 | 4.429 | 1.600 |
| 2005 | 87.540 | 25.051 | 4.105 | 7.727 | 4.195 | 7.881 | 1.759 | 4.559 | 0.700 |
| 2006 | 83.840 | 22.133 | 4.033 | 9.657 | 2.494 | 5.524 | 1.729 | 4.032 | 0.400 |
| 2007 | 96.332 | 15.825 | 3.525 | 12.537 | 1.420 | 6.350 | 1.511 | 3.834 | 0.400 |
| 2008 | 88.290 | 14.940 | 2.934 | 11.085 | 2.567 | 5.496 | 1.258 | 2.740 | 0.231 |
| 2009 | 69.716 | 13.269 | 3.667 | 10.783 | 2.431 | 5.639 | 1.571 | 3.841 | 0.265 |
| 2010 | 70.156 | 17.863 | 3.305 | 9.738 | 1.649 | 5.472 | 1.417 | 3.829 | 0.275 |
| 2011 | 76.744 | 20.447 | 4.176 | 15.505 | 1.543 | 6.013 | 1.790 | 2.914 | 0.185 |
| 2012 | 82.531 | 19.204 | 4.588 | 11.978 | 1.751 | 3.223 | 1.966 | 1.845 | 0.008 |
| 2013 | 101.350 | 23.583 | 5.634 | 14.709 | 2.151 | 3.958 | 2.415 | 2.266 | 0.010 |

Table App.A.2: South and west coast historic (ICSEAF 1989) GLM standardized CPUE data (Glazer, 2013) (the indices in Glazer, 2013, were subsequently updated to include 2012, Glazer, pers. commn) for M. paradoxus and M. capensis.

| Year | GLM CPUE ( $\mathrm{kg} \mathrm{min}^{-1}$ ) <br> Species combined |  | Year | GLM CPUE ( $\mathrm{kg} \mathrm{min}^{-1}$ ) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | M. paradoxus | M. capensis |  |
|  | West Coast | South Coast |  | West Coast | South Coast | West Coast | South Coast |
| 1955 | 17.31 | - |  | 1978 | 4.20 | 1.09 | 0.74 | 1.68 |
| 1956 | 15.64 | - | 1979 | 4.16 | 1.08 | 1.21 | 1.84 |
| 1957 | 16.47 | - | 1980 | 3.86 | 1.59 | 1.08 | 2.07 |
| 1958 | 16.26 | - | 1981 | 3.85 | 1.03 | 1.11 | 2.01 |
| 1959 | 16.26 | - | 1982 | 3.79 | 1.31 | 0.95 | 1.99 |
| 1960 | 17.31 | - | 1983 | 4.08 | 1.43 | 1.28 | 2.46 |
| 1961 | 12.09 | - | 1984 | 4.17 | 1.55 | 1.35 | 2.84 |
| 1962 | 14.18 | - | 1985 | 4.88 | 2.24 | 1.49 | 3.45 |
| 1963 | 13.97 | - | 1986 | 4.25 | 2.18 | 1.22 | 2.87 |
| 1964 | 14.60 | - | 1987 | 3.50 | 1.94 | 1.07 | 2.63 |
| 1965 | 10.84 | - | 1988 | 3.46 | 1.60 | 0.96 | 2.79 |
| 1966 | 10.63 | - | 1989 | 3.74 | 1.57 | 1.03 | 3.00 |
| 1967 | 10.01 | - | 1990 | 4.02 | 2.34 | 0.87 | 3.59 |
| 1968 | 10.01 | - | 1991 | 4.37 | 2.32 | 1.08 | 3.24 |
| 1969 | 8.62 | 1.28 | 1992 | 3.81 | 2.72 | 1.32 | 2.96 |
| 1970 | 7.23 | 1.22 | 1993 | 3.89 | 2.25 | 1.24 | 2.13 |
| 1971 | 7.09 | 1.14 | 1994 | 4.16 | 1.81 | 1.58 | 2.88 |
| 1972 | 4.90 | 0.64 | 1995 | 3.63 | 1.39 | 1.63 | 2.71 |
| 1973 | 4.97 | 0.56 | 1996 | 3.94 | 1.99 | 1.89 | 2.50 |
| 1974 | 4.65 | 0.54 | 1997 | 3.52 | 2.28 | 1.64 | 1.82 |
| 1975 | 4.66 | 0.37 | 1998 | 3.93 | 1.93 | 1.87 | 2.09 |
| 1976 | 5.35 | 0.40 | 1999 | 3.19 | 2.20 | 1.72 | 2.00 |
| 1977 | 4.84 | 0.42 | 2000 | 2.84 | 1.67 | 1.61 | 2.22 |
|  |  |  | 2001 | 2.30 | 1.65 | 1.10 | 1.64 |
|  |  |  | 2002 | 2.23 | 1.41 | 1.06 | 1.84 |
|  |  |  | 2003 | 2.64 | 1.93 | 0.90 | 2.07 |
|  |  |  | 2004 | 2.20 | 1.52 | 0.81 | 1.65 |
|  |  |  | 2005 | 2.19 | 1.34 | 0.58 | 1.48 |
|  |  |  | 2006 | 2.38 | 1.46 | 0.62 | 1.03 |
|  |  |  | 2007 | 2.92 | 1.53 | 0.60 | 0.98 |
|  |  |  | 2008 | 3.23 | 1.65 | 0.81 | 1.56 |
|  |  |  | 2009 | 3.23 | 2.05 | 1.42 | 2.89 |
|  |  |  | 2010 | 3.54 | 2.25 | 1.22 | 2.22 |
|  |  |  | 2011 | 3.30 | 2.62 | 1.47 | 2.79 |
|  |  |  | 2012 | 2.92 | 2.17 | 1.24 | 1.62 |

Table App.A.3: Survey abundance estimates and associated standard errors in thousand tons for M. paradoxus for the depth range $0-500 \mathrm{~m}$ for the South Coast and for the West Coast. Values in bold are for the surveys conducted by the Africana with the new gear. These have not been updated from Fairweather (2012).

| Year | West coast |  |  |  | South coast |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Summer |  | Winter |  | Spring (Sept) |  | Autumn (Apr/May) |  |
|  | Biomass | (s.e.) | Biomass | (s.e.) | Biomass | (s.e.) | Biomass | (s.e.) |
| 1985 | 166.294 | (35.299) | 264.839 | (52.949) | - | - | - | - |
| 1986 | 196.111 | (36.358) | 172.477 | (24.122) | 13.758 | (3.554) | - | - |
| 1987 | 284.805 | (53.101) | 195.482 | (44.415) | 21.554 | (4.605) | - | - |
| 1988 | 158.758 | (27.383) | 233.041 | (64.003) | - | - | 30.316 | (11.104) |
| 1989 | - | - | 468.780 | (124.830) | - | - | - | - |
| 1990 | 282.174 | (78.945) | 226.862 | (46.007) | - | - | - | - |
| 1991 | 327.020 | (82.180) | - | - | - | - | 26.638 | (10.460) |
| 1992 | 226.687 | (32.990) | - | - | - | - | 24.304 | (15.195) |
| 1993 | 334.151 | (50.234) | - | - | - | - | 198.849 | (98.452) |
| 1994 | 330.270 | (58.319) | - | - | - | - | 111.469 | (34.627) |
| 1995 | 324.554 | (80.357) | - | - | - | - | 55.068 | (22.380) |
| 1996 | 430.908 | (80.604) | - | - | - | - | 85.546 | (25.484) |
| 1997 | 569.957 | (108.200) | - | - | - | - | 135.192 | (51.031) |
| 1998 | - | - | - | - | - | - | - | - |
| 1999 | 569.364 | (114.536) | - | - | - | - | 321.478 | (113.557) |
| 2000 | - | - | - | - | - | - | - | - |
| 2001 | - | - | - | - | 19.929 | (9.956) | - | - |
| 2002 | 267.487 | (35.068) | - | - | - | - | - | - |
| 2003 | 411.177 | (69.431) | - | - | 88.442 | (36.051) | 108.857 | (37.528) |
| 2004 | 259.527 | (56.021) | - | - | 63.900 | (17.894) | 48.898 | (20.343) |
| 2005 | 286.416 | (39.849) | - | - | - | - | 26.605 | (7.952) |
| 2006 | 315.310 | (49.490) | - | - | 72.415 | (15.500) | 34.799 | (8.325) |
| 2007 | 397.049 | (71.564) | - | - | 52.287 | (19.231) | 129.646 | (60.661) |
| 2008 | 246.542 | (51.973) | - | - | 24.816 | (8.775) | 39.505 | (11.408) |
| 2009 | 330.235 | (28.526) | - | - | - | - | 102.834 | (28.670) |
| 2010 | 589.533 | (85.686) | - | - | - | - | 169.560 | (67.650) |
| 2011 | 347.082 | (92.540) | - | - | - | - | 24.105 | (7.089) |
| 2012 | 377.515 | (50.690) | - | - | - | - | - | - |
| 2013 | - | - | - | - |  | - | - | - |

Table App.A.4: Survey abundance estimates and associated standard errors in thousand tons for M. capensis for the depth range $0-500 \mathrm{~m}$ for the South Coast and for the West Coast. Values in bold are for the surveys conducted by the Africana with the new gear. These have not been updated from Fairweather (2012).

| Year | West coast |  |  |  | South coast |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Summer |  | Winter |  | Spring (Sept) |  | Autumn (Apr/May) |  |
|  | Biomass | (s.e.) | Biomass | (s.e.) | Biomass | (s.e.) | Biomass | (s.e.) |
| 1985 | 125.028 | (22.719) | 181.487 | (27.476) | - | - | - | - |
| 1986 | 117.810 | (23.636) | 119.587 | (18.489) | 121.197 | (16.625) | - | - |
| 1987 | 75.693 | (10.241) | 87.391 | (11.198) | 159.088 | (17.233) | - | - |
| 1988 | 66.725 | (10.765) | 47.120 | (9.568) | - | - | 165.939 | (21.871) |
| 1989 | - | - | 323.833 | (67.295) | - | - | - | - |
| 1990 | 455.798 | (135.237) | 157.800 | (23.561) | - | - | - | - |
| 1991 | 77.357 | (14.995) | - | - | - | - | 274.298 | (44.395) |
| 1992 | 95.407 | (11.744) | - | - | - | - | 138.085 | (15.357) |
| 1993 | 92.598 | (14.589) | - | - | - | - | 158.340 | (13.733) |
| 1994 | 121.257 | (35.951) | - | - | - | - | 160.555 | (23.701) |
| 1995 | 199.142 | (26.812) | - | - | - | - | 236.025 | (31.840) |
| 1996 | 83.337 | (9.285) | - | - | - | - | 244.410 | (25.107) |
| 1997 | 257.293 | (46.056) | - | - | - | - | 183.087 | (18.906) |
| 1998 | - | - | - | - | - | - | - | - |
| 1999 | 196.992 | (32.059) | - | - | - | - | 191.203 | (14.952) |
| 2000 | - | - | - | - | - | - | - | - |
| 2001 | - | - | - | - | 133.793 | (20.858) | - | - |
| 2002 | 106.253 | (15.813) | - | - | - | - | - | - |
| 2003 | 75.960 | (13.314) | - | - | 82.928 | (9.010) | 128.450 | (20.062) |
| 2004 | 205.939 | (33.216) | - | - | 106.119 | (15.596) | 99.902 | (12.027) |
| 2005 | 70.983 | (13.845) | - | - | - | - | 76.932 | (5.965) |
| 2006 | 88.420 | (22.851) | - | - | 99.867 | (9.803) | 130.900 | (14.816) |
| 2007 | 82.040 | (11.491) | - | - | 74.615 | (7.383) | 70.940 | (5.615) |
| 2008 | 50.877 | (5.355) | - | - | 94.232 | (11.456) | 108.195 | (9.978) |
| 2009 | 175.289 | (39.920) | - | - | - | - | 124.004 | (11.808) |
| 2010 | 163.545 | (34.444) | - | - | - | - | 184.960 | (37.720) |
| 2011 | 89.392 | (23.218) | - | - | - | - | 117.222 | (11.857) |
| 2012 | 92.588 | (11.926) | - | - | - | - | - | - |
| 2013 | - | - | - | - | - | - | - | - |

Table App.A.5: Commercial length frequencies available in November 2013.

|  | Offshore trawl <br> Species combined |  | Inshore trawl <br> M. capensis | Longline Spp combined | Longline <br> . paradoxus | Longline <br> M. capensis | Longline <br> M. capensis |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | WC | SC | SC | WC | WC | WC | SC |
| 1975 | - | $\checkmark$ | - | - | - | - | - |
| 1976 | - | $\checkmark$ | - | - | - | - | - |
| 1977 | - | $\checkmark$ | - | - | - | - | - |
| 1978 | - | $\checkmark$ | - | - | - | - | - |
| 1979 | - | $\checkmark$ | - | - | - | - | - |
| 1980 | - | $\checkmark$ | - | - | - | - | - |
| 1981 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1982 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1983 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1984 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1985 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1986 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1987 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1988 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1989 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1990 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1991 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1992 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1993 | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | - | - |
| 1994 | $\checkmark$ | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | $\checkmark$ |
| 1995 | $\checkmark$ | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | $\checkmark$ |
| 1996 | $\checkmark$ | $\checkmark$ | $\checkmark$ | $\checkmark$ | - | - | $\checkmark$ |
| 1997 | $\checkmark$ | - | $\checkmark$ | $\checkmark$ | - | - | $\checkmark$ |
| 1998 | $\checkmark$ | - | $\checkmark$ | - | - | - | - |
| 1999 | $\checkmark$ | - | $\checkmark$ | - | - | - | - |
| 2000 | - | - | $\checkmark$ | - | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2001 | - | - | - | - | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2002 | - | - | - | - | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2003 | - | - | - | - | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2004 | - | - | - | - | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2005 | $\checkmark$ | - | - | - | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2006 | $\checkmark$ | - | - | - | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2007 | $\checkmark$ | - | $\checkmark$ | - | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2008 | $\checkmark$ | - | $\checkmark$ | - | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2009 | $\checkmark$ | - | $\checkmark$ | - | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2010 | $\checkmark$ | - | $\checkmark$ | - | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2011 | $\checkmark$ | - | $\checkmark$ | - | - | - | - |
| 2012 | $\checkmark$ | - | $\checkmark$ | - | $-$ | $-$ | - |

Table App.A.6: Survey length frequencies available in November 2013.

| Year | West coast |  |  |  | South coast |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Summer |  | Winter |  | Spring (Sept) |  | Autumn (Apr/May) |  |
|  | Sex-aggr. | By sex | Sex-aggr. | By sex | Sex-aggr. | By sex | Sex-aggr. | By sex |
| 1985 | $\checkmark$ | - | $\checkmark$ | - | - | - |  | - |
| 1986 | $\checkmark$ | - | $\checkmark$ | - | $\checkmark$ | - | - | - |
| 1987 | $\checkmark$ | - | $\checkmark$ | - | $\checkmark$ | - | - | - |
| 1988 | $\checkmark$ | - | $\checkmark$ | - | - | - | $\checkmark$ | - |
| 1989 | - | - | $\checkmark$ | - | - | - | - | - |
| 1990 | $\checkmark$ | - | $\checkmark$ | - | - | - | - | - |
| 1991 | $\checkmark$ | - | - | - | - | - | - | - |
| 1992 | $\checkmark$ | - | - | - | - | - | $\checkmark$ | - |
| 1993 | $\checkmark$ | $\checkmark$ | - | - | - | - | $\checkmark$ | $\checkmark$ |
| 1994 | $\checkmark$ | $\checkmark$ | - | - | - | - | $\checkmark$ | $\checkmark$ |
| 1995 | $\checkmark$ | $\checkmark$ | - | - | - | - | $\checkmark$ | $\checkmark$ |
| 1996 | $\checkmark$ | $\checkmark$ | - | - | - | - | $\checkmark$ | $\checkmark$ |
| 1997 | $\checkmark$ | $\checkmark$ | - | - | - | - | $\checkmark$ | $\checkmark$ |
| 1998 | - | - | - | - | - | - | - | - |
| 1999 | $\checkmark$ | $\checkmark$ | - | - | - | - | $\checkmark$ | - |
| 2000 | - | - | - | - | - | - | - | - |
| 2001 | - | - | - | - | $\checkmark$ | - | - | - |
| 2002 | $\checkmark$ | - | - | - | - | - | - | - |
| 2003 | $\checkmark$ | - | - | - | $\checkmark$ | - | $\checkmark$ | - |
| 2004 | $\checkmark$ | - | - | - | $\checkmark$ | - | $\checkmark$ | - |
| 2005 | $\checkmark$ | - | - | - | - | - | $\checkmark$ | - |
| 2006 | $\checkmark$ | $\checkmark$ | - | - | $\checkmark$ | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2007 | $\checkmark$ | $\checkmark$ | - | - | $\checkmark$ | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2008 | $\checkmark$ | $\checkmark$ | - | - | $\checkmark$ | $\checkmark$ | $\checkmark$ | $\checkmark$ |
| 2009 | $\checkmark$ | $\checkmark$ | - | - | - | - | $\checkmark$ | $\checkmark$ |
| 2010 | $\checkmark$ | $\checkmark$ | - | - | - | - | $\checkmark$ | $\checkmark$ |
| 2011 | $\checkmark$ | $\checkmark$ | - | - | - | - | $\checkmark$ | $\checkmark$ |
| 2012 | $\checkmark$ | $\checkmark$ | - | - | - | - | $\checkmark$ | $\checkmark$ |
| 2013 | $-$ | - | $-$ | $-$ | $-$ | $-$ | $-$ | - |

Table App.A.7: Species- and sex-disaggregated age and length data available in November 2013 by reader.


Table App.A.8: Female maturity-at-length ogive parameter estimates (from Singh et al. 2011).

|  | $l_{50}(\mathrm{~cm})$ | $\Delta$ |
| :---: | :---: | :---: |
| $M$. paradoxus | 41.53 | 2.98 |
| $M$. capensis | 53.83 | 10.14 |

Table App.A.9: Length-weight relationship estimates (from Singh 2013).

|  | $\alpha$ <br> $\left(\mathrm{gm} / \mathrm{cm}^{\beta}\right)$ | $\beta$ |
| ---: | :---: | :---: |
| M. paradoxus: |  |  |
| Males | 0.007750 | 2.977 |
| Females | 0.005700 | 3.071 |
| M. capensis: |  |  |
| Males | 0.006750 | 3.044 |
| Females | 0.005950 | 3.075 |

## APPENDIX B: Gender-disaggregated, age-structured production model fitting to age-length keys

The model used is a gender-disaggregated Age-Structured Production Model (ASPM), which is fitted directly to agelength keys (ALKs) and length frequencies. The model also assesses the two species as two independent stocks and is fitted to species-disaggregated data as well as species-combined data. The general specifications and equations of the overall model are set out below, together with some key choices in the implementation of the methodology. Details of the contributions to the log-likelihood function from the different data considered are also given. Quasi-Newton minimisation is used to minimise the total negative log-likelihood function (implemented using AD Model Builder ${ }^{\mathrm{TM}}$, Otter Research, Ltd. (Fournier et al. 2011)).

## App.B. 1 Population Dynamics

## App.B1.. 1 Numbers-at-age

The resource dynamics of the two populations ( $M$. capensis and $M$. paradoxus) of the South African hake are modelled by the following set of equations.
Note: for ease of reading, the 'species' subscript $s$ has been omitted below where equations are identical for the two species.

$$
\begin{align*}
N_{y+1,0}^{g} & =R_{y+1}^{g}  \tag{B.1}\\
N_{y+1, a+1}^{g} & =\left(N_{y a}^{g} e^{-M_{a}^{g} / 2}-\sum_{f} C_{f y a}^{g}\right) e^{-M_{a}^{g} / 2} \quad \text { for } 0 \leq a \leq m-2  \tag{B.2}\\
N_{y+1, m}^{g} & =\left(N_{y, m-1}^{g} e^{-M_{m-1}^{g} / 2}-\sum_{f} C_{f, y, m-1}^{g}\right) e^{-M_{m-1}^{g} / 2} \\
& +\left(N_{y m}^{g} e^{-M_{m}^{g} / 2}-\sum_{f} C_{f y m}^{g}\right) e^{-M_{m}^{g} / 2} \tag{B.3}
\end{align*}
$$

where
$N_{y a}^{g} \quad$ is the number of fish of gender $g$ and age $a$ at the start of year $y^{1}$;
$R_{y}^{g} \quad$ is the recruitment (number of 0 -year-old fish) of fish of gender $g$ at the start of year $y$;
$m \quad$ is the maximum age considered (taken to be a plus-group);
$M_{a}^{g} \quad$ denotes the natural mortality rate on fish of gender $g$ and age $a$; and
$C_{f y a}^{g} \quad$ is the number of hake of gender $g$ and age $a$ caught in year $y$ by fleet $f$.

## App.B.1.2 Recruitment

The number of recruits (i.e. new zero-year old fish) at the start of year $y$ is assumed to be related to the corresponding female spawning stock size (i.e., the biomass of mature female fish). The underlying assumptions are that female spawning output can limit subsequent recruitment, but that there are always sufficient males to provide adequate fertilisation. The recruitment and corresponding female spawning stock size are related by means of the BevertonHolt (Beverton and Holt 1957) or a modified (generalised) form of the Ricker stock-recruitment relationship. These

[^0]forms are parameterized in terms of the "steepness" of the stock-recruitment relationship, $h$, the pre-exploitation equilibrium female spawning biomass, $K^{\circ} s p$, and the pre-exploitation recruitment, $R_{0}$ and assuming a 50:50 sexsplit at recruitment.
$R_{y}^{g}=\frac{4 h R_{0} B_{y}^{\rho, s p}}{K^{\rho, s p}(1-h)+(5 h-1) B_{y}^{\rho, s p}} e^{\left(\varsigma_{y}-\sigma_{R}^{2} / 2\right)}$
for the Beverton-Holt stock-recruitment relationship and
$R_{y}^{g}=\alpha B_{y}^{\odot, s p} \exp \left(-\beta\left(B_{y}^{\odot, s p}\right)^{\gamma}\right) e^{\left(\varsigma_{y}-\sigma_{R}^{2} / 2\right)}$
with
$\alpha=R_{0} \exp \left(\beta\left(K^{\circ, s p}\right)^{\gamma}\right) \quad$ and $\quad \beta=\frac{\ln (5 h)}{\left(K^{\circ, s p}\right)^{\gamma}\left(1-5^{-\gamma}\right)}$
for the modified Ricker relationship (for the true Ricker, $\gamma=1$ ) where
$\varsigma_{y} \quad$ reflects fluctuation about the expected recruitment in year $y$;
$\sigma_{R} \quad$ is the standard deviation of the log-residuals, which is input ( $\sigma_{R}=0.45$ and is taken to decrease from this value to 0.1 over the last five years to statistically stabilise estimates of recent recruitment).
$B_{y}^{\circ}+s p \quad$ is the female spawning biomass at the start of year $y$, computed as:
$B_{y}^{\varrho, s p}=\sum_{a=1}^{m} f_{a}^{\circ} w_{a}^{\circ} N_{y a}^{\text {¢ }}$
where
$w_{a}^{g} \quad$ is the begin-year mass of fish of gender $g$ and age $a$;
$f_{a}^{g} \quad$ is the proportion of fish of gender $g$ and age $a$ that are mature (converted from maturity-at-length, see equation App.B.47); and
$R_{0}=K^{\circ}, s p /\left[\sum_{a=1}^{m-1} f_{a}^{\circ} w_{a}^{\circ} e^{-\sum_{a^{\prime}=0}^{a-1} M_{a}^{g}}+f_{m}^{\circ} w_{m}^{\circ} \frac{e^{-\sum_{a=0}^{m-1} M_{a^{\prime}}^{g}}}{1-e^{-M_{m}^{g}}}\right]$
For the Beverton-Holt form, $h$ is bounded above by 0.98 to preclude high recruitment at extremely low spawning biomass, whereas for the modified Ricker form, $h$ is bounded above by 1.5 to preclude extreme compensatory behaviour.

## App.B.1.3 Total catch and catches-at-age

The fleet-disaggregated catch by mass, in year $y$ is given by:
$C_{f y}=\sum_{g} \sum_{a=0}^{m} \widetilde{w}_{f y, a+1 / 2}^{g} C_{f y a}^{g}=\sum_{g} \sum_{a=0}^{m} \widetilde{w}_{f y, a+1 / 2}^{g} N_{y a}^{g} e^{-M_{a}^{g} / 2} F_{f y} S_{f y a}^{g}$
where
$C_{f y a}^{g}$ is the catch-at-age, i.e. the number of fish of gender $g$ and age $a$, caught in year $y$ by fleet $f$;
$F_{f y} \quad$ is the fishing mortality of a fully selected age class, for fleet $f$ in year $y$ (independent of $g$ );
$S_{f y a}^{g}=\sum_{l} S_{f y l}^{g} P_{a+1 / 2, l}^{g}$
$S_{f y a}^{g} \quad$ is the commercial selectivity of gender $g$ at age $a$ for fleet $f$ and year $y$;
$S_{f y l}^{g} \quad$ is the commercial selectivity of gender $g$ at length / for year $y$, and fleet $f$;
$\widetilde{w}_{f y, a+1 / 2}^{g}=\sum_{l} S_{f y l}^{g} w_{l}^{g} P_{a+1 / 2, l}^{g} / \sum_{l} S_{f y l}^{g} P_{a+1 / 2, l}^{g}$
$\widetilde{w}_{f y, a+1 / 2}^{g}$ is the selectivity-weighted mid-year weight-at-age $a$ of gender $g$ for fleet $f$ and year $y$;
$w_{l}^{g} \quad$ is the weight of fish of gender $g$ and length $/$;
$P_{a+1 / 2, l}^{g}$ is the mid-year proportion of fish of age $a$ and gender $g$ that fall in the length group $/$ (i.e., $\sum_{l} P_{a+1 / 2, l}^{g}=1$ for all ages $a$ ).
The matrix $P$ is calculated under the assumption that length-at-age is log-normally distributed about a mean given by the von Bertalanffy equation, i.e.:
$\ln l_{a} \sim N\left[\ln \left(l_{\infty}\left(1-e^{-\kappa\left(a-t_{0}\right)}\right)\right) ;\left(\frac{\theta_{a}}{l_{\infty}\left(1-e^{-\kappa\left(a-t_{0}\right)}\right)}\right)^{2}\right]$
where $\theta_{a}$ is the standard deviation of length-at-age $a$, which is estimated directly in the model fitting for age 0 , and for ages 1 and above a linear relationship applies:

$$
\theta_{a}=\left\{\begin{array}{cc}
B_{0} & \text { for } a=0 \\
(\beta a+\alpha) & \text { for } 1 \leq a \leq m
\end{array}\right.
$$

with species and gender-specific $B_{0}$, 回 and estimated in the model fitting procedure. A penalty is added so that $\theta_{a}$ is increasing with age, i.e. $\quad \square>0$.

## App.B.1.4 Exploitable and survey biomasses

The model estimate of the mid-year exploitable ("available") component of biomass for each species and fleet is calculated by converting the numbers-at-age into mid-year mass-at-age and applying natural and fishing mortality for half the year:

$$
\begin{equation*}
B_{f y}^{e x}=\sum_{g} \sum_{a=0}^{m} \tilde{w}_{f y, a+1 / 2}^{g} S_{f y a}^{g} N_{y a}^{g} e^{-M_{a}^{g} / 2}\left(1-\sum_{f} S_{f y a}^{g} F_{f y} / 2\right) \tag{B.11}
\end{equation*}
$$

The model estimate of the survey biomass at the start of the year (summer) is given by:

$$
\begin{equation*}
B_{y}^{s u r v}=\sum_{g} \sum_{a=0}^{m_{s}} \widetilde{w}_{a}^{g, s u m} S_{a}^{g, s u m} N_{y a}^{g} \tag{B.12}
\end{equation*}
$$

and in mid-year (winter):

$$
\begin{equation*}
B_{y}^{s u r v}=\sum_{g} \sum_{a=0}^{m} \widetilde{w}_{a+1 / 2}^{g, \text { win }} S_{a}^{g, w i n} N_{y a}^{g} e^{-M_{a}^{g} / 2}\left(1-\sum_{f} S_{f y a}^{g} F_{f y} / 2\right) \tag{B.13}
\end{equation*}
$$

where
$S_{a}^{g, s u m / w i n}$ is the survey selectivity of gender $g$ for age $a$, converted from survey selectivity-at-length in the same manner as for the commercial selectivity (eqn B.8);
$\tilde{w}_{a}^{g, i} \quad$ is the survey selectivity-weighted weight-at-age $a$ of gender $g$ for survey $i$, computed in the same manner as for the commercial selectivity-weight-at-age (equation App.II.9) and taking account of the begin-year ( $\tilde{w}_{y, a}^{g, \text { sum }}$ from $P_{a, l}^{g}$ ) or mid-year $\left(\tilde{w}_{y, a+1 / 2}^{g, \text { win }}\right.$ from $\left.P_{a+1 / 2, l}^{g}\right)$ nature of the surveys.

Note that both the spring and autumn surveys are taken to correspond to winter (mid-year).
It is assumed that the resource is at the deterministic equilibrium that corresponds to an absence of harvesting at the start of the initial year considered, i.e., $B_{1}^{g, s p}=K^{g, s p}$, and the year $y=1$ corresponds to 1917 when catches commence.

## App.B.2 MSY and related quantities

The equilibrium catch for a fully selected fishing proportion $F^{*}$ is calculated as:

$$
\begin{equation*}
C\left(F^{*}\right)=\sum_{g} \sum_{a} \widetilde{w}_{a+1 / 2}^{g} S_{a}^{g} F^{*} N_{a}^{g}\left(F^{*}\right) e^{-\left(\left(M_{a}^{g}+S_{a}^{s} F^{*}\right) / 2\right)} \tag{B.14}
\end{equation*}
$$

where
$S_{a}^{g}$ is the average selectivity across all fleets, for the most recent five years;
$S_{a}^{g}=\frac{\sum_{y=2005}^{2009} \sum_{f} S_{f j a}^{g} F_{f y}}{\max \left(\sum_{y=2005}^{2009} \sum_{f} S_{f j a}^{g} F_{f y}\right)}$
where the maximum is taken over genders and ages; and with

$$
N_{a}^{g}\left(F^{*}\right)=\left\{\begin{array}{cc}
R_{1}\left(F^{*}\right) & \text { for } a=1  \tag{B.16}\\
N_{a-1}^{g}\left(F^{*}\right) e^{-M_{a-1}^{g}}\left(1-S_{a-1}^{g} F^{*}\right) & \text { for } 1<a<m \\
\frac{N_{m-1}^{g}\left(F^{*}\right) e^{-M_{m-1}^{g}}\left(1-S_{m-1}^{g} F^{*}\right)}{\left(1-e^{-M_{m}^{g}}\left(1-S_{m}^{g} F^{*}\right)\right)} & \text { for } a=m
\end{array}\right.
$$

where

$$
\begin{equation*}
R_{1}\left(F^{*}\right)=\frac{\alpha B^{\frac{Q}{, s p}}\left(F^{*}\right)}{\beta+B^{\circ, s p}\left(F^{*}\right)} \tag{B.17}
\end{equation*}
$$

for a Beverton-Holt stock-recruitment relationship.
The maximum of $C\left(F^{*}\right)$ is then found by searching over $F^{*}$ to give $F_{\text {MSY }}^{*}$, with the associated female spawning biomass given by:
$B_{M S Y}^{\odot}{ }^{\circ} s p=\sum_{a} f_{a}^{\circ} w_{a}^{\circ} N_{a}^{\circ}\left(F_{\mathrm{MSY}}^{*}\right)$

## App.B. 3 The likelihood function

The model is fit to CPUE and survey biomass indices, commercial and survey length frequencies, survey age-length keys, as well as to the stock-recruitment curve to estimate model parameters. Contributions by each of these to the negative of the log-likelihood $(-\ell \mathrm{n} L)$ are as follows ${ }^{2}$.

## App.B.3.1 CPUE relative biomass data

The likelihood is calculated by assuming that the observed biomass index (here CPUE) is log-normally distributed about its expected value:
$I_{y}^{i}=\hat{I}_{y}^{i} e^{\varepsilon_{y}^{i}} \quad$ or $\quad \varepsilon_{y}^{i}=\ln \left(I_{y}^{i}\right)-\ln \left(\hat{I}_{y}^{i}\right)$
where
$I_{y}^{i} \quad$ is the biomass index for year $y$ and series $i$ (which corresponds to a specified species and fleet);
$\hat{I}_{y}^{i}=\hat{q}^{i} \hat{B}_{f y}^{e x}$ is the corresponding model estimate, where $\widehat{B}_{f y}^{e x}$ is the model estimate of exploitable resource biomass, given by equation B.11;
$\hat{q}^{i} \quad$ is the constant of proportionality for biomass series $I$; and
$\varepsilon_{y}^{i} \quad$ from $N\left(0,\left(\sigma_{y}^{i}\right)^{2}\right)$.
In cases where the CPUE series are based upon species-aggregated catches (as available pre-1978), the corresponding model estimate is derived by assuming two types of fishing zones: z1) an "M. capensis only zone", corresponding to shallow-water and z2) a "mixed zone" (Figure B.1).

The total catch of hake of both species (BS) by fleet $f$ in year $y\left(C_{B S, f y}\right)$ can be written as:
$C_{B S, f y}=C_{C, f y}^{z 1}+C_{C, f y}^{z 2}+C_{P, f y}$
where
$C_{C, f y}^{z 1} \quad$ is the $M$. capensis catch by fleet $f$ in year $y$ in the $M$. capensis only zone (z1);
$C_{C, f y}^{z 2} \quad$ is the $M$. capensis catch by fleet $f$ in year $y$ in the mixed zone (z2); and
$C_{P, f y} \quad$ is the M. paradoxus catch by fleet $f$ in year $y$ in the mixed zone.
Catch rate is assumed to be proportional to exploitable biomass. Furthermore, let a be the proportion of the $M$. capensis exploitable biomass in the mixed zone ( $\gamma=B_{C, f y}^{e x, z 2} / B_{C, f y}^{e x}$ ) (assumed to be constant throughout the period for simplicity) and $\psi_{f y}$ be the proportion of the effort of fleet $f$ in the mixed zone in year $y\left(\psi_{f y}=E_{f y}^{z 2} / E_{f y}\right)$, so that:
$C_{C, f y}^{z 1}=q_{C}^{i, z 1} B_{C f y}^{e x, z 1} E_{f y}^{z 1}=q_{C}^{i, z 1}(1-\gamma) B_{C, f y}^{e x}\left(1-\psi_{f y}\right) E_{f y}$
$C_{C, f y}^{z 2}=q_{C}^{i, z 2} B_{C, f y}^{e x, z 2} E_{f y}^{z 2}=q_{C}^{i, z 2} \gamma B_{C, f y}^{e x} \psi_{f y} E_{f y}$ and
$C_{P, f y}=q_{P}^{i} B_{P, f y}^{e x} E_{f y}^{z 2}=q_{P}^{i} B_{P, f y}^{e x} \psi_{f y} E_{f y}$
where
$E_{f y}=E_{f y}^{z 1}+E_{f y}^{z 2}$ is the total effort of fleet $f$, corresponding to combined-species CPUE series $i$ which consists of the effort in the $M$. capensis only zone ( $E_{f y}^{z 1}$ ) and the effort in the mixed zone ( $E_{f y}^{z 2}$ );
$q_{C}^{i, z j} \quad$ is the catchability for $M$. capensis $(C)$ for biomass series $i$, and zone $z j$; and

[^1]$q_{P}^{i} \quad$ is the catchability for $M$. paradoxus $(P)$ for biomass series $i$.
It follows that:
$C_{C, f y}=B_{C, f y}^{e x} E_{f y}\left[q_{C}^{i, z 1}(1-\gamma)\left(1-\psi_{f y}\right)+q_{C}^{i, z 2} \gamma \psi_{f y}\right]$
$C_{P, f y}=B_{P, f y}^{e x} E_{f y} q_{P}^{i} \psi_{f y}$
From solving equations B. 23 and B.24:
$s_{f y}=\frac{q_{C}^{i, z 1}(1-\gamma)}{\left\{\frac{C_{C, f y} B_{P, f y}^{e x} q_{P}^{i}}{B_{C, f y}^{e x} C_{P, f y}}-q_{C}^{i, z 2} \gamma+q_{C}^{i, z 1}(1-\gamma)\right\}}$
and:
\[

$$
\begin{equation*}
\hat{I}_{y}^{i}=\frac{C_{f y}}{E_{f y}}=\frac{C_{f y} B_{P, f y}^{e x} q_{P}^{i} \psi_{f y}}{C_{P, f y}} \tag{B.26}
\end{equation*}
$$

\]

| Zone 1 (z1): | Zone 2 (z2): |
| :--- | :--- |
| M. capensis only | Mixed zone |
| M. capensis: | M. capensis: |
| biomass $\left(B_{C}^{z 1}\right), \operatorname{catch}\left(C_{C}^{z 1}\right)$ | biomass $\left(B_{C}^{z 2}\right), \operatorname{catch}\left(C_{C}^{z 2}\right)$ |
|  | M. paradoxus: |
|  | biomass $\left(B_{P}\right), \operatorname{catch}\left(C_{P}\right)$ |
| Effort in zone $1\left(E^{21}\right)$ | Effort in zone $2\left(E^{22}\right)$ |

Figure B.1: Diagrammatic representation of the two conceptual fishing zones.

Two species-aggregated CPUE indices are available: the ICSEAF West Coast and the ICSEAF South Coast series. For consistency, $q$ 's for each species (and zone) are forced to be in the same proportion:

$$
\begin{equation*}
q_{s}^{S C}=r q_{s}^{W C} \tag{B.27}
\end{equation*}
$$

To correct for possible negative bias in estimates of variance $\left(\sigma_{y}^{i}\right)$ and to avoid according unrealistically high precision (and so giving inappropriately high weight) to the CPUE data, lower bounds on the standard deviations of the residuals for the logarithm of the CPUE series have been enforced: for the historic ICSEAF CPUE series (separate West Coast and South Coast series) the lower bound is set to 0.25 , and to 0.15 for the recent GLM-standardised CPUE series, i.e.: $\sigma^{I C S E A F} \geq 0.25$ and $\sigma^{G L M} \geq 0.15$.

The contribution of the CPUE data to the negative of the log-likelihood function (after removal of constants) is then given by:
$-\ell \mathrm{n} L^{\text {CPUE }}=\sum_{i} \sum_{y}\left\lfloor\ln \left(\sigma_{y}^{i}\right)+\left(\varepsilon_{y}^{i}\right)^{2} / 2\left(\sigma_{y}^{i}\right)^{2}\right\rfloor$
where
$\sigma_{y}^{i} \quad$ is the standard deviation of the residuals for the logarithms of index $i$ in year $y$.

Homoscedasticity of residuals for CPUE series is customarily assumed ${ }^{3}$, so that $\sigma_{y}^{i}=\sigma^{i}$ is estimated in the fitting procedure by its maximum likelihood value:

$$
\begin{equation*}
\hat{\sigma}^{i}=\sqrt{1 / n_{i} \sum_{y}\left(\ln \left(I_{y}^{i}\right)-\ell \mathrm{n}\left(\hat{I}_{y}^{i}\right)\right)^{2}} \tag{B.29}
\end{equation*}
$$

where $n_{i}$ is the number of data points for biomass index $i$.
In the case of the species-disaggregated CPUE series, the catchability coefficient $q^{i}$ for biomass index $i$ is estimated by its maximum likelihood value, which in the more general case of heteroscedastic residuals is given by:
$\ln \hat{q}^{i}=\frac{\sum_{y}\left(\ln I_{y}^{i}-\ln \hat{B}_{f y}^{e x}\right) /\left(\sigma_{y}^{i}\right)^{2}}{\sum_{y} 1 /\left(\sigma_{y}^{i}\right)^{2}}$
In the case of the species-combined CPUE, $q_{C}^{W C, z 1}, q_{C}^{W C, z 2}, q_{P}^{W C}, r$ and are estimated directly in the fitting procedure.

## App.B.3.2 Survey biomass data

Data from the research surveys are treated as relative biomass indices in a similar manner to the speciesdisaggregated CPUE series above, with survey selectivity function $S_{a}^{g, \text { sum/win }}$ replacing the commercial selectivity $S_{f y a}^{g}$ (see equations B. 12 and B. 13 above, which also take account of the begin- or mid-year nature of the survey).

An estimate of sampling variance is available for most surveys and the associated $\sigma_{y}^{i}$ is generally taken to be given by the corresponding survey CV. However, these estimates likely fail to include all sources of variability, and unrealistically high precision (low variance and hence high weight) could hence be accorded to these indices. The contribution of the survey data to the negative log-likelihood is of the same form as that of the CPUE biomass data (see equation B.28). The procedure adopted takes into account an additional variance $\left(\sigma_{A}\right)^{2}$ which is treated as another estimable parameter in the minimisation process. This procedure is carried out enforcing the constraint that $\left(\sigma_{A}\right)^{2}>0$, i.e. the overall variance cannot be less than its externally input component.

In June 2003, the trawl gear on the Africana was changed and a different value for the multiplicative bias factor $q$ is taken to apply to the surveys conducted with the new gear. Calibration experiments have been conducted between the Africana with the old gear (hereafter referred to as the "old Africana") and the Nansen, and between the Africana with the new gear ("new Africana") and the Nansen, in order to provide a basis to relate the multiplicative biases of the Africana with the two types of gear ( $q_{\text {old }}$ and $q_{\text {new }}$ ). A GLM analysis assuming negative binomial distributions for the catches made (Brandão et al., 2004) provided the following estimates:

$$
\begin{array}{ll}
\Delta \ell n q^{\text {capensis }}=-0.494 \text { with } \sigma_{\Delta \ell n q^{\text {capensis }}}=0.141 & \text { i.e. }\left(q^{\text {new }} / q^{\text {old }}\right)^{\text {capensis }}=0.610 \quad \text { and } \\
\Delta \ell n q^{\text {paradoxus }}=-0.053 \text { with } \sigma_{\Delta \ell n q^{\text {paradouss }}}=0.117 & \text { i.e. }\left(q^{\text {new }} / q^{\text {old }}\right)^{\text {paradoxus }}=0.948
\end{array}
$$

where

$$
\begin{equation*}
\ell n q_{\text {new }}^{s}=\ln q_{\text {old }}^{s}+\Delta \ell n q^{s} \quad \text { with } s=\text { capensis or paradoxus } \tag{B.31}
\end{equation*}
$$

No plausible explanation has yet been found for the particularly large extent to which catch efficiency for $M$. capensis is estimated to have decreased for the new research survey trawl net. It was therefore recommended (BENEFIT 2004)

[^2]that the ratio of the catchability of the new to the previous Africana net be below 1 , but not as low as the ratio estimated from the calibration experiments. $\Delta \ell n q^{\text {capensis }}$ is therefore taken as -0.223 , i.e. $\left(q^{\text {new }} / q^{\text {old }}\right)^{\text {capensis }}=0.8$.
The following contribution is therefore added as a penalty (or a log prior in a Bayesian context) to the negative loglikelihood in the assessment:
$-\ell n L^{q-c h}=\left(\ell n q_{\text {new }}-\ell n q_{\text {old }}-\Delta \ell n q\right)^{2} / 2 \sigma_{\Delta \ell n q}^{2}$
A different length-specific selectivity is estimated for the "old Africana" and the "new Africana".
The survey's coefficients of catchability $q$ (for the survey with the old Africana gear) are constrained to values below 1 (i.e. it is assumed that the nets do not herd the hake):
\[

$$
\begin{equation*}
\text { pen }^{q}=\sum_{i}\left(q_{o l d}^{i}-1\right)^{2} / 0.02^{2} \quad \text { if } \quad q_{\text {old }}^{i}>1 \tag{B.33}
\end{equation*}
$$

\]

## App.B.3.3 Commercial proportions at length

Commercial proportions at length cannot be disaggregated by species and gender. The model is therefore fit to the proportions at length as determined for both species and gender combined.

The catches at length are computed as:
$C_{f y l}=\sum_{s} \sum_{g} \sum_{a=0}^{m} N_{s y a}^{g} F_{s f y} S_{s f y l}^{g} P_{s, a+1 / 2, l}^{g} e^{-M_{s a}^{g} / 2}\left(1-\sum_{f} S_{s f y a}^{g} F_{f y} / 2\right)$
with the predicted proportions at length:
$\widehat{p}_{y l}^{i}=C_{f y l} / \sum_{l^{\prime}} C_{f y l}$.
The contribution of the proportion at length data to the negative of the log-likelihood function when assuming an "adjusted" lognormal error distribution is given by:
$\left.-\ell \mathrm{n} L^{\mathrm{length}}=0.1 \sum_{y} \sum_{l} \ln \left(\sigma_{l e n}^{i} / \sqrt{p_{y l}^{i}}\right)+p_{y l}^{i}\left(\ln p_{y l}^{i}-\ln \hat{p}_{y l}^{i}\right)^{2} / 2\left(\sigma_{l e n}^{i}\right)^{2}\right\rfloor$
where
the superscript ' $i$ ' refers to a particular series of proportions at length data which reflect a specified fleet, and species (or combination thereof); and
$\sigma_{\text {len }}^{i} \quad$ is the standard deviation associated with the proportion at length data, which is estimated in the fitting procedure by:

$$
\begin{equation*}
\hat{\sigma}_{l e n}^{i}=\sqrt{\sum_{y} \sum_{l} p_{y l}^{i}\left(\ln p_{y l}^{i}-\ln \hat{p}_{y l}^{i}\right)^{2} / \sum_{y} \sum_{l} 1} \tag{B.37}
\end{equation*}
$$

The initial 0.1 multiplicative factor is a somewhat arbitrary downweighting to allow for correlation between proportions in adjacent length groups. The coarse basis for this adjustment is the ratio of effective number of ageclasses present to the number of length groups in the minimisation, under the argument that independence in variability is likely to be more closely related to the former.

Commercial proportions at length are incorporated in the likelihood function using equation B.36, for which the summation over length $I$ is taken from length $I_{\text {minus }}$ (considered as a minus group) to $I_{\text {plus }}$ (a plus group). The length for the minus- and plus-groups are fleet specific and are chosen so that typically a few percent, but no more, of the fish sampled fall into these two groups.

## App.B.3.4 Survey proportions at length

The survey proportions at length are incorporated into the negative of the log-likelihood in an analogous manner to the commercial catches-at-age, assuming an adjusted log-normal error distribution (equation B.36). In this case however, data are disaggregated by species, and for some surveys further disaggregated by gender:
$p_{s y l}^{g, \text { surv }}=\frac{C_{s y l}^{g, \text { surv }}}{\sum_{g} \sum_{l^{\prime}} C_{s y l^{\prime}}^{g, \text { surv }}}$ is the observed proportion of fish of species $s$, gender $g$ and length $/$ from survey surv in year $y$; and
$\hat{p}_{s y l}^{g, \text { surv }}$ is the expected proportion of fish of species $s$, gender $g$ and length / in year $y$ in the survey surv, given by:
$\hat{p}_{s y l}^{g, s u r v}=\frac{\sum_{a} S_{s l}^{g, s u m} P_{\text {sal }}^{g} N_{s y a}^{g}}{\sum_{g} \sum_{l^{\prime}} \sum_{a} S_{s l^{\prime}}^{g, s u m} P_{s a l}^{g} N_{\text {sya }}^{g}}$
for begin-year (summer) surveys, or
$\hat{p}_{s y l}^{g, s u r v}=\frac{\sum_{a} S_{s l}^{g, w i n} P_{s, a+1 / 2, l}^{g} N_{s y a}^{g} e^{-M_{s a}^{g} / 2}\left(1-\sum_{f} S_{s f j a}^{g} F_{s f y} / 2\right)}{\sum_{g} \sum_{l^{\prime}} \sum_{a} S_{s l^{\prime}}^{g, w i n} P_{s, a+1 / 2, l}^{g} N_{s y a}^{g} e^{-M_{s a}^{g} / 2}\left(1-\sum_{f} S_{s f j a}^{g} F_{s f y} / 2\right)}$
for mid-year (autumn, winter or spring) surveys.

## 8.II.3.5 Age-length keys

Under the assumption that fish are sampled randomly with respect to age within each length-class, the contribution to the negative log-likelihood for the ALK data (ignoring constants) is:
$-\ln L^{A L K}=-w \sum_{i} \sum_{l} \sum_{a}\left[A_{i, l, a}^{\text {obs }} \ln \left(\hat{A}_{i, l, a}\right)-A_{i, l, a}^{\text {obs }} \ln \left(A_{i, l, a}^{\text {obs }}\right)\right]$
where
$w \quad$ is a downweighting factor to allow for overdispersion in these data compared to the expectation for a multinomial distribution with independent data; this weight factor is somewhat arbitrarily set to 0.01 to avoid these data overriding trend information in the indices of biomass;
$A_{i, a, l}^{\text {obs }}$ is the observed number of fish of size class / that fall in age $a$, for ALK $i$ (a specific combination of survey, year, species and gender);
$\hat{A}_{i, a, l}$ is the model estimate of $A_{i, a, l}^{\text {obs }}$, computed as:

$$
\begin{equation*}
\hat{A}_{i, a, l}=W_{i, l} \frac{C_{i, a, l} \tilde{A}_{a, l}}{\sum_{a^{\prime}} C_{i, a^{\prime}, l} \widetilde{A}_{a^{\prime}, l}} \tag{B.41}
\end{equation*}
$$

where
$W_{i, l} \quad$ is the number of fish in length class / that were aged for ALK $i$,
$\tilde{A}_{a, l}=\sum_{a} P\left(a^{\prime} \mid a\right) A_{a, l}$ is the ALK for age $a$ and length / after accounting for age-reading error,
with $P\left(a^{\prime} \mid a\right)$, the age-reading error matrix, representing the probability of an animal of true age $a$ being aged to be that age or some other age $a^{\prime}$.

Age-reading error matrices have been computed for each reader and for each species as reported in Appendix 8.III.
When multiple readers age the same fish, these data are considered to be independent information in the model fitting.

## App.B.3.6 Stock-recruitment function residuals

The stock-recruitment residuals are assumed to be log-normally distributed. Thus, the contribution of the recruitment residuals to the negative of the log-likelihood function is given by:
$-\ell n L^{S R}=\sum_{s}\left[\sum_{y=y 1}^{y 2} \varsigma_{s y}{ }^{2} / 2 \sigma_{R}^{2}+\left(\sum_{y=y 1}^{y 2} \varsigma_{s y}\right)^{2} / 0.01^{2}\right]$
where
$\varsigma_{s y} \quad$ is the recruitment residual for species $s$, and year $y$, which is assumed to be log-normally distributed with standard deviation $\sigma_{R}$ and which is estimated for year $y 1$ to $y 2$ (see equation B.4) (estimating the stock-recruitment residuals is made possible by the availability of catch-at-age data, which give some indication of the age-structure of the population); and
$\sigma_{R} \quad$ is the standard deviation of the log-residuals, which is input.
The stock-recruitment residuals are estimated for years 1985 to 2006, with recruitment for other years being set deterministically (i.e. exactly as given by the estimated stock-recruitment curve) as there is insufficient catch-at-age information to allow reliable residual estimation for earlier years. A limit on the recent recruitment fluctuations is set by having the $]_{R}$ (which measures the extent of variability in recruitment - see equation - App.II.42) decreasing linearly from 0.45 in 2004 to 0.1 in 2009, effectively forcing recruitment over the last years to lie closer to the stockrecruitment relationship curve.

The second term on the right hand side is introduced to force the average of the residuals estimated over the period from $y 1$ to $y 2$ to be close to zero, for reasons elaborated in the main text.

## App.B. 4 Model parameters

## App.B.4.1 Estimable parameters

The primary parameters estimated are the species-specific female virgin spawning biomass $\left(K_{s}^{q_{s p}}\right)$ and "steepness" of the stock-recruitment relationship $\left(h_{s}\right)$. The standard deviations $\sigma^{i}$ for the CPUE series residuals (the speciescombined as well as the GLM-standardised series) as well as the additional variance $\left(\sigma_{A}^{i}\right)^{2}$ for each survey biomass series are treated as estimable parameters in the minimisation process. Similarly, in the case of the species-combined CPUE, $q_{C}^{W C, z 1}, q_{C}^{W C, z 2}, q_{P}^{W C}, r$ and are directly estimated in the fitting procedure.

The species- and gender-specific von Bertalanffy growth curve parameters ( $I_{\infty}, \kappa$ and $t_{0}$ ) are estimated directly in the model fitting process, as well as $B_{0}$, 国园 and $[$, values used to compute the standard deviation of the length-at-age $a$.

The following parameters are also estimated in the model fits undertaken (if not specifically indicated as fixed):

## App.B.4.1.1 Natural mortality:

Natural mortality $\left(M_{s a}^{g}\right)$ is assumed to be age-specific and is calculated using the following functional form (the selection of the specific form here is based on convenience and is somewhat arbitrary):
$M_{s a}^{\circ}=\left\{\begin{array}{lll}M_{s 2}^{\circ} & \text { for } & a \leq 1 \\ \alpha_{s}^{M}+\frac{\beta_{s}^{M}}{a+1} & \text { for } & 2 \leq a \leq 5 \\ M_{s 5}^{\circ} & \text { for } & a>5\end{array}\right.$
and

$$
\begin{equation*}
M_{s a}^{\text {males }}=v^{s} M_{s a}^{\text {females }} \tag{B.44}
\end{equation*}
$$

$M_{s 0}$ and $M_{s 1}$ are set equal to $M_{s 2}\left(=\alpha_{s}^{M}+\beta_{s}^{M} / 3\right)$ as there are no data (hake of ages younger than 2 are rare in catch and survey data) which would allow independent estimation of $M_{s 0}$ and $M_{s 1}$.

When $M$ values are estimated in the fit, a penalty is added to the total $-\operatorname{lnL}$ so that $M_{s 2} \geq M_{s 5}$ :

$$
\begin{equation*}
\operatorname{pen}^{M}=\sum_{s}\left(M_{s 5}-M_{s 2}\right)^{2} / 0.01^{2} \quad \text { if } M_{s 2}<M_{s 5} \tag{B.45}
\end{equation*}
$$

## App.B.4.1.2 Stock-recruitment residuals:

Stock-recruitment residuals $\varsigma_{s y}$ are estimable parameters in the model fitting process. They are estimated separately for each species from 1985 to the present, and set to zero pre-1985 because there are no catch-at-length data for that period to provide the information necessary to inform estimation.

Table B. 1 summarises the estimable parameters, excluding the selectivity parameters.

## App.B.4.1.3 Survey fishing selectivity-at-length:

The survey selectivities are estimated directly for seven pre-determined lengths for M. paradoxus and M. capensis. When the model was fitted to proportion-at-age rather than proportion-at-length, survey selectivities were estimated directly for each age (i.e. seven age classes). The lengths at which selectivity is estimated directly are survey specific (at constant intervals between the minus and plus groups) and are given in Table B.2. Between these lengths, selectivity is assumed to change linearly. The slope from lengths $I_{\text {minus }}$ to $I_{\text {minus }}+1$ is assumed to continue exponentially to lower lengths down to length 1 , and similarly the slope from lengths $I_{\text {plus }}-1$ to $I_{\text {plus }}$ for $M$. paradoxus and $M$. capensis to continue for greater lengths.

For the South Coast spring and autumn surveys, gender-specific selectivities are estimated for M. paradoxus. Furthermore, the female selectivities are scaled down by a parameter estimated for each of these two surveys to allow for the male predominance in the survey catch.

A penalty is added to the total -InL to smooth the selectivities to smooth the selectivities by penalising deviations from straight line dependence (the choice of a weighting of 3 was made empirically to balance this term having sufficient but not undue influence) :

$$
\begin{equation*}
\text { pen }^{\text {survS }}=\sum_{i} \sum_{L=L_{1}+1}^{L_{7}-1} 3\left(S_{L-1}^{i}-2 S_{L}^{i}+S_{L+1}^{i}\right)^{2} \tag{B.46}
\end{equation*}
$$

where $i$ is a combination of survey, species and gender.

## App.B.4.1.4 Commercial fishing selectivity-at-length:

The fishing selectivity-at-length (gender independent) for each species and fleet, $S_{s f l}$, is estimated in terms of a logistic curve given by:
$S_{s f l}=\left[1+\exp \left(-\left(l-l_{s f}^{c}\right) / \delta_{s f}^{c}\right)\right]^{-1}$
where
$l_{s f}^{c} \mathrm{~cm}$ is the length-at-50\% selectivity,
$\delta_{s f}^{c} \mathrm{~cm}^{-1}$ defines the steepness of the ascending limb of the selectivity curve.
The selectivity is sometimes modified to include a decrease in selectivity at larger lengths, as follows:
$S_{s f l}=S_{s, f, l-1} e^{-s_{s f}} \quad$ for $1>I_{\text {slope }}$,
where
$S_{\text {sfl }}$ measures the rate of decrease in selectivity with length for fish longer than $I_{\text {slope }}$ for the fleet concerned, and is referred to as the "selectivity slope"; and
$I_{\text {slope }}$ is fixed externally from the model, values for each fleet and species are given in Table B.2.
Periods of fixed and changing selectivity have been assumed for the offshore trawl fleet to take account of the change in the selectivity at low ages over time in the commercial catches, likely due to the phasing out of the (illegal) use of net liners to enhance catch rates.

On the South Coast, for M. paradoxus, the female offshore trawl selectivity (only the trawl fleet is assumed to catch M. paradoxus on the South Coast) is scaled down by a factor taken as the average of those estimated for the South Coast spring and autumn surveys. Although there is no gender information for the commercial catches, the South Coast spring and autumn surveys catch a much higher proportion of male $M$. paradoxus than female (ratios of about 7:1 and 3.5:1 for spring and autumn respectively). This is assumed to reflect a difference in distribution of the two genders which would therefore affect the commercial fleet similarly.

Details of the fishing selectivities (including the number of parameters estimated) that are used in the assessment are shown in Table B.4.

## App.B.4.2 Input parameters and other choice for application to hake

## App.B.4.2.1 Age-at-maturity:

The proportion of fish of species $s$, gender $g$ and length / that are mature is assumed to follow a logistic curve with the parameter values given below (from Singh et al. 2011)):

Table B.1: Female maturity-at-length logistic curve parameter values for the new Reference Case.

|  | $l_{50}(\mathrm{~cm})$ | $\Delta$ |
| ---: | :---: | :---: |
| $M$. | 41.53 | 2.98 |
| paradoxus |  |  |
| M. capensis | 53.83 | 10.14 |

Maturity-at-length is then converted to maturity-at-age as follows:

$$
\begin{equation*}
f_{s a}^{g}=\sum_{l} f_{s l}^{g} P_{a, l}^{g} \tag{B.49}
\end{equation*}
$$

## App.B.4.2.2 Weight-at-length:

The weight-at-length for each species and gender is calculated from the mass-at-length function, with values of the parameters for this function listed below (from Fairweather 2008, taking the average of the West and South coasts):

Table B.2: Weight-at-length parameter values.

|  | $\alpha$ <br> $\left(\mathrm{gm} / \mathrm{cm}^{\beta}\right)$ | $\beta$ |
| ---: | :---: | :---: |
| M. paradoxus: |  |  |
| Males | 0.00775 | 2.977 |
| Females | 0.00570 | 3.071 |
| M. capensis: |  |  |
| Males | 0.00675 | 3.044 |
| Females | 0.00595 | 3.075 |

## App.B.4.2.3 Minus- and plus-groups

Because of a combination of gear selectivity and mortality, a relatively small number of fish in the smallest and largest length classes are caught. In consequence, there can be relatively larger errors (in terms of variance) associated with these data. To reduce this effect, the assessment is conducted with minus-and plus-groups obtained by summing the data over the lengths below and above $I_{\text {minus }}$ and $I_{\text {plus }}$ respectively. The minus- and plus-group used are given in Table B.5. Furthermore, the proportions at length data (both commercial and survey) are summed into 2 cm length classes for the model fitting.

Table B.1: Parameters estimated in the model fitting procedure, excluding selectivity parameters.

|  | No of parameters | Parameters estimated |
| :---: | :---: | :---: |
| $K^{9}$ | 2 | $\ln \left(K_{\text {cap }}\right)$ and $\ln \left(K^{\text {¢ }}\right.$ para $)$ |
| $h$ | 2 | $h_{\text {cap }}$ and $h_{\text {para }}$ |
| $M_{a}$ | $4(6)^{*}$ | For each species: $M_{2}, M_{5}($ and $v$ ) |
| Additional variance | 2 | $\sigma_{A, \text { cap }}$ and $\sigma_{A, p a r a}$ |
| Recruitment residuals | 50 | $\zeta_{\text {cap,1985-2009 }}$ and $\zeta_{\text {para,1985-2009 }}$ |
| $\sigma_{\text {CPUE }}$ | 6 | 1 for each series (lower bounds imposed) |
| ICSEAF CPUE | 5 | $q_{\mathrm{C}}{ }^{W C, z i}, q_{\mathrm{C}}{ }^{\mathrm{WC}, z 2}, q_{\mathrm{P}}{ }^{\mathrm{WC}}, r \text { and } \gamma$ |
| $\theta_{a}$ | 12 | For each species and gender: $\theta_{0}, \theta_{1}$ and $\theta_{14}$ |
| Growth | 12 | For each species and gender: $L_{5}, x$ and $t_{0}$ |

Table B.2: Lengths (in cm ) at which survey selectivity is estimated directly.

| $\begin{aligned} & \text { n } \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & \text { si } \end{aligned}$ | West coast summer | 13 | 18 | 23 | 28 | 32 | 37 | 42 | 47 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | West coast winter | 13 | 18 | 24 | 29 | 35 | 40 | 46 | 51 |
|  | South coast spring | 21 | 26 | 30 | 35 | 39 | 44 | 48 | 53 |
|  | South coast autumn | 21 | 26 | 31 | 36 | 42 | 47 | 52 | 65 |
| $\begin{aligned} & \text { n } \\ & \text { U } \\ & \text { S } \\ & \text { ¿ } \end{aligned}$ | West coast summer | 13 | 20 | 26 | 33 | 39 | 46 | 52 | 59 |
|  | West coast winter | 13 | 17 | 21 | 30 | 40 | 47 | 54 | 61 |
|  | South coast spring | 13 | 19 | 28 | 38 | 46 | 54 | 63 | 71 |
|  | South coast autumn | 13 | 19 | 28 | 36 | 44 | 52 | 61 | 69 |

Table B.3: Length $(\mathrm{cm})$ at which selectivity starts to decrease $\left(l_{\text {slope }}\right)$ for each species and fleet.

|  | M. paradoxus | M. capensis |
| :--- | :---: | :---: |
| WC offshore trawl | 40 | 70 |
| SC offshore trawl | 70 | 70 |
| SC inshore trawl | - | 55 |
| WC longline | 85 | 85 |
| SC longline | - | 85 |
| SC handline | - | 70 |

Table B.4: Details for the commercial selectivity-at-length for each fleet and species combination for the new RC, as well as indications of what data are available.

|  | No of est. parameters | M. paradoxus <br> Comments | No of est. parameters | M. capensis <br> Comments | data available |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1. West coast offshore <br> 1917-1976 <br> 1977-1984 <br> 1985-1992 <br> 1993-2013 | 2 <br> 0 <br> 3 | set equal to 1989 <br> two logistic parameters estimated (same slope as 1993+) <br> linear change between 1984 and 1993 <br> selectivity <br> two logistic + slope parameters estimated | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | set equal to 1989 <br> differential shift compared to 1993+ as for paradoxus, slope $1 / 3$ of inshore linear change between 1984 and 1993 selectivity <br> same as SC inshore but shifted to the right by 5 cm , slope $1 / 3$ of inshore | species combined <br> species combined <br> species combined |
| 2. South coast offshore <br> 1917-1976 <br> 1977-1984 <br> 1985-1992 <br> 1993-2013 | 0 <br> 0 <br> 0 <br> 3 <br> 0 | set equal to 1989 <br> differential shift compared to 1993+ as for WC (same slope as 1993+) <br> linear change between 1984 and 1993 <br> selectivity <br> two logistic + slope parameters estimated <br> female downscaling factor (av. of SC spring and autumn surveys's factors) | 0 <br> 0 <br> 0 <br> 0 | set equal to 1989 <br> differential shift compared to 1993+ as for paradoxus, slope $1 / 3$ of inshore linear change between 1984 and 1993 selectivity same as SC inshore but shifted to the right by 5 cm , slope $1 / 3$ of inshore | species combined <br> species combined <br> species combined |
| 3. South coast inshore | - | - | 3 | two logistic + slope parameters estimated | M. capensis |
| 4. West coast longline <br> 1984-1999 <br> 2000-2005 <br> 2006-2013 | $\begin{aligned} & 3 \\ & 1 \\ & 1 \end{aligned}$ | two logistic + slope parameters estimated <br> shift of the ascending limb shift of the ascending limb | $\begin{aligned} & 3 \\ & 1 \\ & 1 \end{aligned}$ | two logistic + slope parameters estimated <br> shift of the ascending limb shift of the ascending limb | species combined <br> species disaggregated <br> species disaggregated |
| 5. South coast longline <br> 1984-1999 <br> 2000-2005 <br> 2006-2013 |  |  | $\begin{aligned} & 3 \\ & 1 \\ & 1 \end{aligned}$ | two logistic + slope parameters estimated <br> shift of the ascending limb shift of the ascending limb | species combined species disaggregated species disaggregated |
| 6. South coast handline | - | - | 0 | parameters taken as average of SC longline and inshore parameters |  |
| West coast summer survey <br> Africana old <br> Africana new <br> West coast winter survey <br> Africana old <br> South coast spring survey <br> Africana old <br> Africana new <br> South coast autumn <br> Africana old <br> Africana new | 7 <br> 5 <br> 7 <br> 7 <br> 5 <br> 1 <br> 7 <br> 5 <br> 1 | estimated for 7 specified lengths same slope as old estimated for 7 specified lengths estimated for 7 specified lengths same slope as old female downscaling factor estimated for 7 specified lengths same slope as old female downscaling factor | 7 5 <br> 7 <br> 7 <br> 5 <br> 7 <br> 5 | estimated for 7 specified lengths same slope as old estimated for 7 specified lengths estimated for 7 specified lengths same slope as old estimated for 7 specified lengths same slope as old | species disaggregated species disaggregated <br> species disaggregated <br> species disaggregated species disaggregated <br> species disaggregated <br> species disaggregated |
| Total | 58 |  | 56 |  |  |

Table B.5: Minus- and plus-groups taken for the surveys and commercial proportion at length data.

| SURVEY DATA | M. paradoxus |  |  | M. capensis |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
|  | Minus | Plus | Minus | Plus |  |
| West coast summer | 13 | 47 | 13 | 59 |  |
| West coast winter | 13 | 51 | 13 | 61 |  |
| South coast spring | 21 | 53 | 13 | 71 |  |
| South coast autumn | 21 | 65 | 13 | 69 |  |

## COMMERCIAL DATA

|  | Minus | Plus |
| :--- | :---: | :---: |
| West coast offshore, species combined | 23 | 65 |
| South coast offshore, species combined | 27 | 75 |
| South coast inshore, $M$. capensis | 27 | 65 |
| West coast longline, species combined | 51 | 91 |
| South coast longline, $M$. capensis | 51 | 91 |
| Both coasts offshore, species combined | 25 | 65 |

## Appendix C - Base Case fit to Age-Length Keys

## ALK1 <br> est coast summe

1990







ALK7
West coast summer

1996













Fig. App.B.1a: Observed vs predicted mean age-at-length for M. paradoxus males and females for the west coast summer surveys.


Fig. App.B.1b: Observed vs predicted mean age-at-length for M. paradoxus males and females for the west coast winter and south coast spring surveys.

ALK21
ALK21
South coast autum
1992
ALK22
outh coast autumn
1993
ALK23
outh coast autumn
1994
ALK24
South coast autumn
1995
ALK25
South coast autumn
1996











ALK30
South coast autumn
ALK31
South coast autumn









Fig. App.B.1c: Observed vs predicted mean age-at-length for M. paradoxus males and females for the south coast autumn surveys.


Fig. App.B.1d: Observed vs predicted mean age-at-length for M. paradoxus males and females for the commercial offshore trawl and commercial longline.

## FISHERIES/2013/NOV/SWG-DEM67

## MARAM IWS/DEC13/Hake/P2





$$
\begin{gathered}
\text { ALK4 } \\
\text { West coast summer }
\end{gathered}
$$

$$
\begin{gathered}
\text { ALK5 } \\
\text { West coast summer }
\end{gathered}
$$


ALK7 West coast summer
1993
0
$\vdots$
$\vdots$
$\vdots$














ALK12
West coast summer
ALK13
West coast summer
ALK14
West coast summer
2008








Fig. App.B.2a: Observed vs predicted mean age-at-length for $M$. capensis males and females for the west coast summer surveys.
ALK15
West coast winter
1988



ALK17
South coast spring
2006
ALK18
South coast spring
2007
ALK19

$$
\begin{aligned}
& \text { South coast spring } \\
& 2008
\end{aligned}
$$






Fig. App.B.2b: Observed vs predicted mean age-at-length for $M$. capensis males and females for the west coast winter and south coast spring surveys.

## FISHERIES/2013/NOV/SWG-DEM67

## MARAM IWS/DEC13/Hake/P2

ALK20
South coast aut
1991


$$
\begin{gathered}
\text { ALK21 } \\
\text { South coast autumn } \\
1992
\end{gathered}
$$



$$
\begin{gathered}
\text { ALK24 } \\
\text { South coast autumn } \\
1994
\end{gathered}
$$



$$
\begin{gathered}
\text { ALK26 } \\
\text { South coast autum } \\
1996
\end{gathered}
$$









South coast autum
1997
ALK28
South coast autumn
1999
South coast autumn
1999







Fig. App.B.2c: Observed vs predicted mean age-at-length for $M$. capensis males and females for the south coast autumn surveys.


Fig. App.B.2d: Observed vs predicted mean age-at-length for $M$. capensis males and females for the commercial offshore trawl and commercial longline.


[^0]:    ${ }^{1}$ In the interests of less cumbersome notation, subscripts have been separated by commas only when this is necessary for clarity.

[^1]:    ${ }^{2}$ Strictly it is a penalised log-likelihood which is maximised in the fitting process, as some contributions that would correspond to priors in a Bayesian estimation process are added.

[^2]:    ${ }^{3}$ There are insufficient data in any series to enable this to be tested with meaningful power.

