**The size-structured (length-based) stock assessment methodology applied to west coast rock lobster**

S.J. Johnston and D.S. Butterworth

MARAM

Department of Mathematics and Applied Mathematics

University of Cape Town

Rondebosch

Cape Town

**A The population dynamics model**

The model of the west coast rock lobster dynamics is described by a size-structured model which operates at a 1mm interval scale, from *l* = 1mm to *l* = *lmax*, where *l* refers to the carapace length (which for brevity will also be referred to as “size”) in mm. The largest size class in the model, *lmax*, is equal to 200mm. Male and female lobsters are modelled separately, with ***m*** referring to males and ***f*** referring to females. In this report, the superscript ***s*** refers to either the male or female sex, where the two sexes are treated separately but similarly. Although the fishing season starts towards the end of year *t* and runs into year *t*+1, it is denoted as season *t* (unless otherwise specified). Where data are collected and represented in 5mm size class intervals, the size class *x* refers to lobsters of sizes *x*, *x*+1, *x*+2 …*x*+4 mm.

### **A.1 Basic dynamics**

The basic model in vector-matrix notation is:

  (A1.1)

 is the (1 x 200) vector of number of males/females in each of the 1mm size

 classes at the start of season *t,*

 is the (200 x 200) somatic growth matrix for males/females over season *t* (models the proportion of lobsters moving from one size class to another as a result of somatic growth),

 is the (200 x 200) fishing survivorship matrix for males/females over season *t,*

 is the (200 x 200) diagonal natural survivorship matrix for males/females (constant over time), and

 is the (1 x 200) recruitment vector for season *t*+1 (the same for males and for females).

Note that the structure of this discrete model is such that harvest is followed by natural mortality and then by somatic growth, i.e. fishing is approximated by a pulse at the start of the season.

### **A.2 Length-weight conversions**

The model operates on a numbers basis, but the number of lobsters in each size class is easily converted to biomass using the length-weight conversion equations as follows:

 **** for males, and (A2.1)

  for females (A2.2)

where the total body weight, *w*, is in grams, and length *l*, is the carapace length in centimetres.

***A.3 Equilibrium starting conditions***

Assuming no harvesting (pristine conditions), the equilibrium populations numbers are calculated as follows:

  (A3.1)

and therefore

  (A3.2)

 represents the male/female pristine equilibrium population vector, and *I* is the unit matrix.

Equation (A3.2) is used to set up the initial pristine population size-structure in 1910 when catches are assumed to have commenced.

### **A.4 Natural survivorship matrix S**

The natural survivorship matrix  models the proportion of lobsters in each size class at the beginning of the season which survive to the beginning of the next season in the absence of fishing. This matrix consists of zeros except for the diagonal entries which contain the values for size class *l*. The natural survivorship is assumed to be independent of size for lobsters greater than or equal to 60mm. However, the demography of immature lobsters is likely to be somewhat different from that of mature lobsters, so that smaller lobsters are assumed to have reduced survivorship. A linear increase in survivorship from 0.1 at size *l* = 1mm (an artificial size used purely for definition convenience) to (males) and (females) at size 60mm is assumed. The maximum male annual natural survivorship, , is set at 0.90 for the Base Case. The maximum female natural survivorship, , is an estimable parameter of the model. Natural survivorship is assumed to be constant over time. The specific entries on the diagonal of the natural survivorship matrix are calculated as follows.

(Alternate values for male survivorship will be examined in robustness tests).

Males:

  for mm (A4.1)

  for mm

Females:

  for mm (A4.2)

  for mm

### **A.5 Gear selectivit**y

Different sexes and sizes of lobsters have different probabilities of being caught by various fishing gear, and each gear type operates in different parts of the overall habitat area where different sizes or sexes may aggregate at different times. The fishing selectivity functions model these different probabilities of being caught by the gear. In this model, selectivity is thus really a combination of the physical gear selectivity as well as availability.

For the Trap, Hoopnet and FIMS selectivity functions (for both male and female), the function for each gear is calculated assuming linear interpolation between a number of estimable parameters for various carapace lengths (usually 5mm categroes):

For male and female trap selectivity we estimate:

the selectivity of a lobster carapace length *l* and of sex *s* (*m* or *f*)*,*

for CL 62mm, 67mm, 72mm, 77mm, 82mm, 87mm, 92mm, 97mm and 102+mm. Selectivity is assumed zero at CL 45mm.

For male and female hoop selectivity we estimate:

the selectivity of a lobster carapace length *l* and of sex *s* (*m* or *f*)*,*

for CL 72mm, 77mm, 82mm, 87mm, 92mm, 97mm and 102+mm. Selectivity is assumed zero at CL 45mm.

For male FIMS selectivity we estimate:

the selectivity of a lobster carapace length *l* and of sex *m,*

for CL 57mm, 62mm, 67mm, 72mm, 77mm, 82mm, 87mm, 92mm. Selectivity is assumed zero at CL 45mm and 100mm.

For female FIMS selectivity we estimate:

the selectivity of a lobster carapace length *l* and of sex *f,*

for CL 57mm, 62mm, 67mm, 72mm, 77mm. Selectivity is assumed zero at CL 45mm and 100mm.

Comment: In many cases we are getting a strong downward trend as length increases (see IWS/DEC10/WCRLA/P3). Is this realistic? What could be the mechanism. Should we be forcing the RHS of the selectivity curve to be flat?

***A.6 Fishing survivorship matrix H***

The fishing survivorship matrix *H* consists of diagonal entries  with zeros everywhere else, where the  terms, the proportion of lobsters in size class *l* caught during season *t* for the specified catch

type , are calculated as follows:

  (A6.1)

where *c* refers to either the commercial trap or hoop, poaching or recreational catch.

For trap and hoops:

  for 

  for 

where  is the commercial legal minimum size limit in season *t*.

For poaching:

  for all *l*

For recreational:

  for 

  for 

 where is the recreational minimum size limit

The parameter *d* refers to the **discard mortality**, and is the proportion of undersized lobsters that are caught and returned to the sea which subsequently die as a result of the original capture-and-return process. A value of ***d* = 0.10** is used.

The factor  in equation A6.1 is the fishing mortality for fully selected animals, which links the model-generated catches to the observed catches by weight (catches are only recorded by weight, not number), and is calculated as follows:

 , (A6.2)

where

 *Ctc* is the observed catch (trap, hoop, poaching or recreational) by weight in season *t*, and

  and  are the numbers-at-length for season *t*.

Note: for poaching, the  is summed over all sizes, not just above the minimum size.

### **A.7 The 1991 fishing season**

The 1991 season in the model receives special treatment as the minimum size was altered during this season and only males were harvested. After this alteration, both males and females were however subject to discard mortality. The model thus calculates the  and  in two steps. The first refers to the initial part of the season when the minimum size limit was 89mm (denoted by superscript 89), while the second refers to the subsequent part when the minimum size limit was 75mm (denoted by superscript 75). Therefore, for the initial part:

  (A7.1)

where

  for *l <* 89mm for males

  for females of all sizes

  for *l ≥* 89mm for males

and

 

The fully-selected fishing mortality  above is calculated as follows:

  (A7.2)

where  refers to the catch (by weight) taken in the initial part of the season when the minimum size limit was 89mm. Before proceeding with the calculations for the second part of the season, the numbers-at-size are updated to the end of the first part of the season taking into account the losses due to fishing mortality only, i.e.:

 , and also therefore (A7.3)

 

where

 , and



For the subsequent part of the season:

  (A7.4)

where

  for *l <* 75mm for males

  for females of all sizes

  for *l ≥* 75mm for males

and

 

where, further:

  (A7.5)

and  is the catch (by weight) taken in the subsequent part of the season to which a minimum size limit of 75mm applied,

 , and



We therefore have , and (A7.6)

 .

### **A.8 Somatic growth projection matrix A**

The somatic growth projection matrix models the proportion of lobsters in one size class that move into another size class between seasons as a result of somatic growth. The entry  represents the proportion of lobsters in size class *i* that move into size class *j* during the moult at the end of season *t*. Moult increments of lobsters have been shown to be highly variable. This variance in somatic growth is modelled using a normal probability distribution truncated at ± 3 standard deviations. In order to calculate the  values using this probability distribution, the mean annual moult increment value, , must be evaluated as well as an estimate of standard deviation around this mean moult increment, , for each length *l* and year *t.* The calculation of the  and  values are reported in detail in the Appendix. Note that this distribution allows for the possibility that lobsters shrink during the moult.

The calculation of the  values are described by the following steps for any given season *t* and sex *s* - the *t* and *s* indices have been omitted below for less cluttered notation.

[Note that because *i* represents the length class (*i-*1*, i*) mm, e.g. *i* = 75 means animals of size *l* where 74 < *l*  75 mm, it is convenient to use the midpoint *i-*0.5 of this interval in the following calculations.]

Step 1: For any given *i*, initialise *a( i,j )* to zero for *j* = 1, 2 ,...200, and calculate the mean length after moult (*mi*) for length *li* (where *li* = *i* - 0.5):

  (A8.1)

where  is the mean moult increment for length *i*.

Step 2: Calculate the two endpoints of the growth distribution:

  (lower point) (A8.2)

  (upper point)

where  is the standard deviation for the mean moult increment *gi*.

If , then set *a(i,j)* = 1 [where *j* is such that ] and go back to step 1, incrementing *i* by 1.

Step 3: Find the midpoints *mp*(1) and *mp*(2) of the length classes into which *mi*(1) and *mi*(2) respectively fall.

Step 4: Define *mpj* as the midpoint corresponding to length class ( *j*-1, *j*); then for

 *mpj* = *mp*(1)*, mp*(1)*+*1*, ..., mp*(2) do the following.

A: If *mpj = mp*(1), set *y*(1) *= mi*(1)(A8.3)

 set *y*(2) *= mpj* + 0.5 (A8.4)

 If *mpj = mp*(2), set *y*(1) *= mpj* - 0.5 (A8.5)

 set *y*(2) *= mi*(2)(A8.6)

 For all other cases, set *y*(1) *= mpj* - 0.5 (A8.7)

 set *y*(2) *= mpj* + 0.5 (A8.8)

B: Calculate the *z* values for *y*(1) and *y*(2) (to convert to a normal distribution N(0;1)):

  and (A8.9)

  (A8.10)

C: Calculate the area under the normal curve for each of these z values:

*A*(1) = area from *z*(1) to 0

*A*(2) = area from *z*(2) to 0

D: Calculate the proportion of lobsters moving from *i* to *j* (equal to the area under the normal curve from *z*(1) to *z*(2)).

If *1 < mpj* < 199, do the following:

 *j* = *mpj* + 0.5

  if *y*(1) *< mi < y*(2)(A8.11)

 otherwise

 

If *mpj* < 1, do the following:

 *j* = 1

  (A8.12)

If *mpj* > 199, do the following:

 *j* = 200

  (A8.13)

Step 5: Normalise *a(i,j)* so that 

This is to correct for the fact that normal distributions are used at Step 4B, whereas Step 2 explicitly limits growth to  three standard deviations from the mean.

Step 6: Go back to Step 1, incrementing *i* by 1.

***A.9 Recruitment R***

Recruitment is assumed to contribute to the first 15 size classes only. The value of recruitment for each of these size classes in season *t* is *R(t)*. Recruitment is assumed equal for male and female lobsters, and hence:

  for mm (A9.1)

  for *l >* 15mm

Recruitment is assumed to be linear between *R* (1910), *R* (1920), *R* (1950), *R* (1970), *R* (1975), *R* (1980), *R* (1985), *R* (1990),

*R* (1995), *R*((2000) and  *R*((2003). *R*(2007)+ is calculated as the geometric mean of the six values *R*(1975)-*R*(2000).

***A.10 Catch***

The model is constrained so that the model-predicted catch exactly equals the “observed” catch . The “observed” catch consists of three contributions in each season t:

 i) the recorded commercial catch,  = 

 ii) an estimate of poaching, , and

 iii) an estimate of the public recreational catch, .

Thus:

  (A10.1)

The model-predicted catch is calculated by:

  (A10.2)

##### A10.1 Commercial catches

Although the commercial fishery is known to have began around 1870, the model assumes the resource is at pristine state in 1910, but does take into account the small catches estimated to have been made between 1870 and 1910 - see Fisheries/2010/SEP/SWG-WCRL/22 for further details.

*A.10.2 Poaching and recreational catches*

Poaching:

Poaching (at a non-negligible level) is assumed to start in 1951 and increase linearly to a value *Pmax* in the year 1990 (these assumptions are based on anecdotal reports from the industry). Poaching is assumed to remain at the *Pmax* level after 1990. The estimates of used in the model are thus:

  for *t* < 1950 (A10.3)

  for 

  for *t* > 1990

The reference case assessment assumes *Pmax* = 500 MT based upon rough suppositions by SFRI personnel for the area-aggregated model.

Recreational Catch:

Recreational catches (at a non-negligible level) are assumed to have started in 1959 (this assumption was based on discussions held by the RLWG in conjunction with the industry). From 1959 to 1992 the recreational catches are assumed to have increased linearly from zero to 469 MT (as estimated in the telephone survey for that year). Telephone survey estimated values are used for seasons 1993+. The estimates of  (in tons) used in the model are reported in FISHERIES/2010/AUG/SWG-WCRL19.

Interim Relief

Interim relief estimates for the resource as a whole are available for the 2007-2009 seasons. These catches would be added to the commercial hoopnet catches for each area for the relevant season.

### **A.11 Biomass**

Since the model maintains information on numbers-at-size by sex, the biomass of any desired component of the resource is readily computed. These biomass estimates can refer to the male portion of the stock, , the female portion of the stock, , or to the combined total of males and females *B*. Furthermore, a biomass estimate can be calculated for any given size range. For example, the biomass estimate for male and female lobsters above size *l* = 75mm in season *t* is calculated as follows:

  (A11.1)

All biomass estimates quoted in this paper have units metric tons (MT) or ‘000 MT.

***A.12 MSY Calculations***

The maximum sustainable yield (MSY) can be calculated for a number of scenarios. The following must be defined:

 i) somatic growth rate,

 ii) recruitment level,

 iii) gear selectivity, and

 iv) minimum legal size.

Equations A.3.4 and A.3.5 are modified to include the fishing survivorship matrix, so that an equilibrium size structure can be calculated for any level of harvesting. For a fishing mortality level *F*, which defines a fishing survival matrix *H*:

  (A12.1)

so that

  (A12.2)

The sustainable yield for any *F* can be calculated as follows:

 SY(*F*) =  + )) (A12.3)

where the  are the number of male/female lobsters in the equilibrium population at size *l* and at fishing mortality level *F.* At some level of fishing mortality, the sustainable yield reaches a maximum - the maximum sustainable yield (MSY). The *F* value corresponding to MSY is denoted *F*MSY.

The biomass level (above the legal minimum size) at which MSY occurs, *B*MSY, is calculated as follows:

 *B*MSY =  + ) (A12.4)

###### **A.13 The Likelihood function**

The estimable parameters of the size-structured model are estimated by fitting the model to a number of data series. This is achieved by minimising the negative of the log of the likelihood function for these data.

A.13.1 Data included in the likelihood

The negative log likelihood function incorporates three types of data.

*i) CPUE*

Three sources of CPUE data for the resource are used.

1. Commercial trap CPUE data for the 1981-2008 period.
2. Commercial hoopnet CPUE data for the 1981-2008 period.
3. FIMS CPUE data for the 1992-2008 period.

(Remember, 1995 refers to the 1995/96 season, for example.)

These CPUE values are assumed to be log-normally distributed about their model-based expectations.

*ii) Catch-at-size data*

Four sources of catch-at-size data are used.

1. Data obtained from commercial trap samples for 1976-2008.
2. Data obtained from commercial hoopnet samples for 1976-2008.
3. Data obtained from the FIMS samples for 1992-2008.
4. Data obtained from trap samples which cover the full size range including the sub-legal size classes for 1994-1998.

Males and females are treated separately. Data are reported by 5mm size classes, i.e. the *n*th size class incorporates lobsters of size *n* mm, *n*+1 mm ... *n*+4 mm. Data for size classes equal or larger than the minimum legal size at the time are used from the commercial trap and hoopnet data. Data for the full size range are used for both the FIMS and the trap sub-legal data. Females are omitted for the 1991 season as there was a ban on catching females for that season. Due to there being very few observations for some size classes, plus- and minus-groups have been formed (indicated in bold) where necessary. This ensures that no observed size-class has a percentage less than 1%.

*iii) Percent females in the catch (F%)*

Three sources of data are used.

 a) Data from the commercial trap samples used for calculating catch-at-size data for 1976-2008.

1. Data from the commercial hoopnet samples used for calculating catch-at-size data for 1976-2008.

 c) Data from the FIMS samples for 1992-2008.

Percent females (*F*%) values are the percentage by **number** of females in the samples. Once again the season 1991 is omitted as there was a ban on catching females then.

A.13.2 The Likelihood function

The superscript *s* is sometimes used below. It refers to either of the two sexes, where males and females are treated separately but identically. After removal of constants from the negative log likelihood, the function to be minimised is:

 

  (A13.1)

 + -lnL (F%)

 + *Rpen*

[Note that the catch-at-size data are downweighted by a factor of 0.1.]

where

*A*  is the data series identifier:

where *A* = 1 refers to the commercial trap data

 *A* = 2 refers to the commercial hoopnet data

 *A* = 3 refers to the FIMS data

 *A* = 4 refers to the trap sublegal catch-at-size data

*t* is the season

*l* is the length class in 5 mm intervals

 is the numbers of seasons of CPUE data for series *A*

 is the observed CPUE value in season *t* for series *A*

 is the model-estimated CPUE value in season *t* for series *A*

 is the observed catch-at-size proportion for male/female lobsters of size *l* in season *t* for series *A*

 is the model-estimated catch-at-size proportion for male/female lobsters of size *l* in season *t* for series *A*

 is the observed percentage females in the catch in season *t* for series *A*

 is the model-estimated percentage females in the catch in season *t* for series *A*

Also:

 .100 (A13.2)

  (A13.3)

   (A13.4)

   (A13.5)

For series *A* = 1 and *A* = 2: (trap and hoopnet data)

 (A13.6)

 (A13.7)

For series *A* = 3 : (FIMS data)

 (A13.8)

 (A13.9)

The reference case model has the following penalty function added to the –*ln*L (this reflects “shrinkage to the mean”. or in Bayesian terms using a prior that reflects the recent past distribution of recruitments):

 (A13.10)

where

 (A13.11)

 (A13.12)

**F% likelihood function**

The method of including the F% (percentage females in the catch) data in the likelihood function was changed when the assessments were expanded to include area-disaggregated approaches. This change was necessary, as previously all the area-aggregated observed F% data were greater than zero. Moving to the super-area level, it was found that there were some instances where the true observed F% for a given year was in fact zero. The then current likelihood function required taking logarithms of these observed values, which produced a mathematical problem (one can’t take a logarithm of zero!). It was thus decided to move to a binomial probability distributional assumption for the F% likelihood function calculation. The F% likelihood contribution is now calculated as follows:



where

*N* is the sample size

is the observed female **proportion** by number in catch in year *y*

is the model estimated female **proportion by number** in catch in year *y*

To ensure one does not have a computational problem with taking logarithms of a zero value, the following also applies (based on the fact that the limit as  of  is 0):

If =0

 X=0

else

 X=

and

If =1

 Y=0

else

 Y=ln(1-)

where

 (A13.10)

**Calculation of the sample size *N***

We do not have the various sample sizes, but we require an estimate of *N* for the above –ln*L* equation. What is done is to calculate an approximation to the effective sample size based on the model fit as follows:

For each F% series calculate *N*, so that . (A13.11)

This will be done separately for each super-area using the 2009 updated assessment model.

**F%  values**

The F% **** values are calculated as follows:

 (A13.12)

**Reference case and alternate Models**

The two alternate models (Alt1 and Alt2) are virtually identical to the RC model, except with regards to the *R*2003 value. For the RC model *R*2003  is an estimable parameter, (although previously it was found to be estimated with very low precision e.g. for Area 8 the 95% CI was 0.0001-1.65), and so is dominated in the estimation by the contribution from equation (1). For this reason, Alt1 and Alt2 models would correspond almost exactly to the RC best fit parameter values except for *R*2003 which would be fixed at the (approximate) upper and lower 25%iles of this distribution as follows:

 (A13.13)

and

 (A13.14)

where  is from equation (2\*) above, and the  value (0.727) corresponds to the 25%iles of a *t*-distribution with the appropriate number of degrees of freedom.

# Appendix: Somatic growth model

In order to compute the entries of the projection matrix, in the size-structured population model (see section A.8), estimates of the mean annual growth (moult) increment values for each size class are needed. Values are calculated separately for males () and females (). Adult lobsters (70mm carapace length and larger) are assumed to moult once a year. Juveniles are known to moult several times a year, so that both the moult increment and moult frequency need to be taken into account when calculating the annual moult increment values for juveniles.

#### ***App1.1 Mature males***

The annual moult increment for a male lobster length *l* for year *t*, is modelled:

  (App1.1)

See RLWS/DEC05/DAT/6/1/3/3 and RLWS/DEC05/DAT/6/1/3/4 for details on the GLM. The GLM assumes that the slope parameter (*ρ*) remains constant over time and that all growth variation can be modelled by a parallel shift of the growth curve each year, thus changing the growth axis-intercept value *β*. Linear interpolation is used ofr years where data are missing.

#### ***App1.2 Mature females***

Moult increment data for female lobsters are scarce, and limited to two periods only: 1969-1971 (Newman and Pollock 1977) and 1992 (SFRI data, unpublished). For 1969, data are available from Dassen Island, Cape Point and Robben Island. For 1992, data are available for Cape Point and Hout Bay. As for males, a GLM approach was used to calculate a  value for each of the two data periods. The slope parameter is estimated to be *ρ* = - 0.0206. As with males:

  (App1.2.1)

In order to calculate the values for the years for which no data are available, linear interpolation is used where the female  values are assumed to follow the same interannual fluctuating pattern as calculated for the male  values.

***App1.3 Juveniles***

Juvenile annual moult increments refer to lobsters smaller than 60mm. Male and female lobsters are assumed to have equal juvenile moult increments. Unlike mature lobsters which are assumed to moult only once a year, juvenile lobsters moult several times a year, with the moult frequency declining gradually with increasing size. The annual moult increment calculation for a juvenile lobster must therefore take the moult frequency or inter-moult period, as well as the inter-moult growth increment, into account.

*App1.3.1. The juvenile inter-moult period*

Only limitted juvenile moult frequency data are available (Pollock 1973). A linear model is fitted to this data:

  (App1.3.1.1)

where is the inter-moult period in days for length *l* (mm), when fitted to these data yields

 *α* = 73.70 days, and

 *β* = 1.36 days.mm-1 .

*App1.3.2. The juvenile moult increment model*

Data are available for three periods: 1973, 1993 and 1995.The quadratic growth model for juveniles for each period is:

 for 1973 *l*  60 (App1.3.2.1)

 for 1993 *l*< 15mm

 for 1993 mm

 for 1993 mm

 for 1995 *l*< 15mm

 for 1995 mm

 for 1995 mm

where

*l* = the mean length before moult (mm)

= the mean moult increment (mm) for a juvenile lobster in season *t*.

, and where .

The model above was fitted to all three sets of juvenile moult increment data – see Table 1 for estimated parameter values. The net **annual** growth for juvenile lobsters is thus:

  for 5mm < *l* 60mm (App1.3.2.2)

Further, to avoid negative growth for very small lobsters, it is assumed that  for mm.

*App1.3.3 Interpolation between adults and juveniles*

A straight line is fitted to the corresponding sets of relative 70mm male moult increments and the  and , and the  and values to calculate both  and  respectively in years for which there is no data, and hence no estimates of  and .

The equations to be used for calculating  and  for all years are:

for :

 (App1.3.3.1)

where $δ\_{1}^{t}=X^{t}^{β}$

 is the relative 70mm male moult increment in year *t* = 

$$β=\frac{ln⁡(0.734)}{ln⁡(0.625)}$$

and similarly, for :

$δ\_{2}^{t}=X^{t}^{β}$ (App1.3.3.2)

where
$$β=\frac{ln⁡(0.715)}{ln⁡(0.625)}$$

***App1.4.Moult increment variance estimate****s*

The projection matrix calculations of require not only estimates of the mean annual moult increment for each length, but also estimates of the associated variances around these values. Standard deviations for moult increments have been calculated from the moult increment data. Moult increment variance models for male, female and juvenile lobsters are as follows. The variances for both adult male and female moult increments are considered to vary with the moult increment value.

##### App1.4.1 Mature males

 for  mm (App1.4.1.1)

 for  mm (App1.4.1.2)

where,

*g* is the moult increment value

*l*  is the length before moult









(See Butterworth *et a.l* 1999 for details).

##### App1.4.2 Mature females

 (App1.4.2.1)

where,

mm,



and

*X* the male average over the 21 years of growth data (1968-1998) calculated to be 1.88mm.

The female growth variances follow a similar but damped trend as for males for varying values of *g*. The *l* dependence for males is lesser, and marked only for larger lengths (110mm) seldom reached by females, so is neglected in this proposal.

##### App1.4.3 Juveniles

There is a standard deviation associated with

 i) the inter-moult period *SD(p)*, and

 ii) with the inter-moult growth increment *SD(Δg)*.

The  (a constant) was estimated from the linear model fit to Pollock’s (1973) data. The  was estimated in the process of fitting the quadratic function to the available data. The estimated values are as follows:

  days

  mm

  mm-1

If equation (App1.3.2.2) is rewritten with simplified notation:

 

then, using a delta method of approximation, the coefficient of variation (CV) of *g* is given by:

  (App1.4.3.1)

where

 

and

 

so that

   (App1.4.3.2)

Note: Unlike the variance for adult lobster growth, the juvenile variance is time-dependent (as Δ*g* is time-dependent).

##### App1.4.4 Interpolation between juveniles and adults

As with the annual moult increment values, linear interpolation is used to provide the standard deviation estimates between 60mm (juvenile) and 70mm (adult) for male lobsters, and between 50mm (juvenile) and 70mm (adult) for females.

#### ***References***

Butterworth, D.S., Glazer, J.P. and S.J.Johnston. 1999. Length and expected growth dependence of moult increment variance estimates for mature male lobster. MCM internal document, WG/10/99/WCL42.

Glazer, J. 1999. GLM analysis applied to the west coast rock lobster growth data. MCM internal document. WG/07/99/WCL22.

Hazell, A. and D. Schoeman (1999) Revisiting juvnile growth of *Jasus lalandii*. MCM internal document. WG/09/99/WCL29.

Newman,G.G. and D.E.Pollock 1977. The population dynamics of rock lobster *Jasus lalandii* on two South African fishing grounds. *Colln Sci. Pap. Int. Commn. SE. Atl. Fish. 4*: 175-187.

Pollock, D.E. 1973. Growth of juvenile rock lobster *Jasus lalandii*. *Investl. Rep. Sea Fish. Branch S.Afr*. 106: 1-16.

Table 1: Best fit parameter values for the juvenile moult increment model.

|  |  |
| --- | --- |
| **Parameter** | **Best fit value** |
|  | -0.439 mm |
|  | 0.2345 |
|  | -0.0027 mm-1 |
|  | 0.265 mm |
|  | 0.0044 mm-1 |
|  | 0.845 |
|  | 0.785 |
|  | 0.714 |
|  | 0.705 |